# Determination of diagnostic species with statistical fidelity measures 

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#### Abstract

Statistical measures of fidelity, i.e. the concentration of species occurrences in vegetation units, are reviewed and compared. The focus is on measures suitable for categorical data which are based on observed species frequencies within a vegetation unit compared with the frequencies expected under random distribution. Particular attention is paid to Bruelheide's $u$ value. It is shown that its original form, based on binomial distribution, is an asymmetric measure of fidelity of a species to a vegetation unit which tends to assign comparatively high fidelity values to rare species. Here, a hypergeometric form of $u$ is introduced which is a symmetric measure of the joint fidelity of species to a vegetation unit and vice versa. It is also shown that another form of the binomial $u$ value may be defined which measures the asymmetric fidelity of a vegetation unit to a species. These $u$ values are compared with phi coefficient, chi-square, $G$ statistic and Fisher's exact test. Contrary to the other measures, phi coefficient is independent of the number of relevés in the data set, and like the hypergeometric form of $u$ and the chi-square it is little affected by the relative size of the vegetation unit. It is therefore particularly useful when comparing species fidelity values among differently sized data sets and vegetation units. However, unlike the other measures it does not measure any statistical significance and may produce unreliable results for small vegetation units and small data sets. The above measures, all based on the comparison of observed/expected frequencies, are compared with the categorical form of the Dufrêne-Legendre Indicator Value Index, an index strongly underweighting the fidelity of rare species.

These fidelity measures are applied to a data set of 15989 relevés of Czech herbaceous vegetation. In a small subset of this data set which simulates a phytosociological table, we demonstrate that traditional table analysis fails to determine diagnostic species of general validity in different habitats and large areas. On the other hand, we show that fidelity calculations used in conjunction with large data sets can replace expert knowledge in the determination of generally valid diagnostic species. Averaging positive fidelity values for all species within a vegetation unit is a useful approach to measure quality of delimination of the vegetation unit. We propose a new way of ordering species in synoptic species-by-relevé tables, using fidelity calculations.


Keywords: Character species; Differential species; Frequency; Indicator species; Non-parametric statistics; Phytosociological database; Plant community; Vegetation classification.

Nomenclature: Ehrendorfer (1973).

## Introduction

The diagnostic species is an important concept in vegetation classification (Whittaker 1962; Westhoff \& van der Maarel 1973). Diagnostic species include species which preferably occur in a single vegetation unit (character species) or in a few vegetation units (differential species). Most frequently, diagnostic species are considered a posteriori, i.e. vegetation samples (relevés) are first classified by expert judgement or a numerical method, and then species with the highest concentration in particular vegetation units are determined as diagnostic. Diagnostic species are useful for identification of previously distinguished vegetation units in field surveys.

In European phytosociology, the concept of diagnostic species has been associated with fidelity, which is a measure of species concentration in vegetation units. However, hardly any attempt has been made to develop and apply a statistical measure of fidelity since Szafer \& Pawłowski (1927) published a quantitative (yet intuitive and not statistical) guide to fidelity determination, which was taken over by Braun-Blanquet (1928) and also by some recent reviews and textbooks of vegetation classification (Westhoff \& van der Maarel 1973; Dierschke 1994). Another intuitive approach was proposed by Bergmeier et al. (1990) who defined diagnostic (character) species as those whose frequency in the vegetation unit was at least two times higher and two frequency classes higher than in the other vegetation units. Generally, the intuitive approaches failed to find a balance between the frequency proportions and frequency differences within and outside the vegetation units and ignored the number of relevés in vegetation units and in the total data set (Barkman 1989).

A statistical measure related to fidelity was developed by Brisse et al. (1995). They used information on species co-occurrences (called fidelities in their terminology) in a large database to define 'species discrimi-
nation capability' (or 'discriminant power') for a vegetation unit. However, this measure evaluates each species on the basis of its co-occurrences with other species rather than according to its actual occurrence within and outside the vegetation unit. As a result, even a species absent from a particular vegetation unit can have a high positive discrimination capability for the vegetation unit. This approach is similar to the Index of Sociological Favourability proposed by Beals (1984) but it is not a direct measure of fidelity.

In this paper, we focus on fidelity measures suitable for categorical data. In vegetation data sets, species presences/absences give a more robust fidelity estimation than covers/abundances as they are less affected by temporal fluctuations and observer bias. In addition, arbitrary weighting of species abundances is avoided by using categorical data (Bruelheide 2000). A promising approach to fidelity measuring is comparison of observed frequencies of species occurrence in the vegetation unit with the frequencies that would be expected if the species distribution was random. This is the theoretical basis of the fidelity measure $u$, developed by Bruelheide $(1995,2000)$. Besides $u$ there are more traditional methods which can be used to measure fidelity, e.g. the chi-square statistic (Goodall 1953; Juhász-Nagy 1964) or $G$ statistic of the likelihood ratio test (also called $G$ test or 2I test; Botta-Dukát \& Borhidi 1999). Except for the implementation of the chi-square statistic in the TABORD program (van der Maarel et al. 1978), the use of statistical fidelity measures was limited to pilot studies and did not become widespread in vegetation science. Recently, Dufrêne \& Legendre (1997) proposed an Indicator Value Index which is also suitable for measuring fidelity and is available in the PCORD package (McCune \& Mefford 1999).

Determination of diagnostic species does not only depend on the appropriate fidelity measure but also on the data set structure. Usually, diagnostic species are determined in data sets which include only relevés of a single alliance, order or class, or of a few closely related vegetation units. Such diagnostic species may have a low potential for generalization, as species affinities to other vegetation units not included in the data set are ignored. With the availability of large phytosociological databases in recent years (Hennekens \& Schaminée 2001), diagnostic species of more general validity should be preferably determined in data sets that include relevés of most vegetation types occurring in a wide area.

The objectives of this paper are (1) reviewing statistical methods suitable for measuring fidelity with categorical data; (2) defining relationships between different fidelity measures and testing their performance in data sets of varying structure; (3) discussing applications of statistical fidelity measures for determining
diagnostic species, evaluating vegetation units and ordering species in classified relevé tables.

## $u$-values and other statistical measures of fidelity

## Variables and contingency tables

We will use the same notation as Bruelheide (1995, 2000):
$N=$ number of relevés in the data set;
$N_{p}=$ number of relevés in the particular vegetation unit;
$n=$ number of occurrences of the species in the data set;
$n_{p}=$ number of occurrences of the species in the particular vegetation unit.

Each relevé in which the species is present is counted as an occurrence of the species. We compare the observed distribution of these occurrences within the data set to what would theoretically be expected if such occurrences were distributed randomly. Observed frequencies can be summarized in the following $2 \times 2$ contingency table. They will be further referred to as $f(o)_{i}$, where $i=1,2,3$, and 4 for the four fields of the table:

| Number of relevés ... | in the <br> vegetation unit | not in the <br> vegetation unit |
| :--- | :---: | :---: |
| containing the species | $n_{p}$ | $n-n_{p}$ |
| not containing the species | $N_{p}-n_{p}$ | $N-N_{p}-n+n_{p}$ |

Corresponding expected frequencies, further referred to as $f(e)_{i}$ are:

| Number of relevés ... | in the <br> vegetation unit | not in the <br> vegetation unit |
| :--- | :---: | :---: |
| containing the species | $n \cdot N_{p} / N$ | $n \cdot\left(N-N_{p}\right) / N$ |
| not containing the species | $(N-n) \cdot N_{p} / N$ | $(N-n) \cdot\left(N-N_{p}\right) / N$ |

u values
Bruelheide $(1995,2000)$ proposed the fidelity measure $u$, which compares the observed number of occurrences of the species in the vegetation unit $\left(n_{p}\right)$ with the expected number of occurrences $\left(\mu=n \cdot N_{p} / N\right)$. The $u$ value is defined as the deviation of the observed frequency of the species occurrence in the vegetation unit from the expected frequency, compared with the standard deviation ( $\sigma$ ):
$u=\left(n_{p}-\mu\right) / \sigma$
In the $2 \times 2$ contingency table above, assuming that $N, N_{p}$ and $n$ are fixed quantities, which is the case in databases, the random variable $n_{p}$ will have a hypergeometric distribution if occurrence of the species is independent of the vegetation unit. Therefore, we use
the standard deviation for a hypergeometric random variable (Sokal \& Rohlf 1995: 94):
$\sigma_{\text {hyp }}=\sqrt{n \cdot N_{p} \cdot(N-n) \cdot\left(N-N_{p}\right) /\left(N^{2} \cdot(N-1)\right)}$
We denote the corresponding $u$ value as $u_{h y p}$. If the species and the vegetation unit co-occur more often than is expected, $u_{\text {hyp }}$ will be positive and will indicate some degree of joint fidelity of the species to the vegetation unit and of the vegetation unit to the species (JuhászNagy 1964; Botta-Dukát \& Borhidi 1999).

Bruelheide $(1995,2000)$ used a binomial approach. For a binomial random variable,
$\sigma_{b i n}=\sqrt{K \cdot P \cdot(1-P)}$
where $K$ is the number of trials and $P$ is the probability of success (Sokal \& Rohlf 1995: 77).

In Bruelheide's approach, a trial is a relevé containing the species in question. There are $n$ such relevés. Success is the event that the releve belongs to the vegetation unit in question, and the probability of success is $P=N_{p} / N$. The assumption of this approach is that $n$ and the relative size of the vegetation unit within the data set $\left(N_{p} / N\right)$ are fixed quantities. Thus the standard deviation is computed as follows:

$$
\begin{equation*}
\sigma_{b i n B}=\sqrt{n \cdot\left(N_{p} / N\right) \cdot\left(1-N_{p} / N\right)} \tag{4}
\end{equation*}
$$

where the subscript $\operatorname{bin} B$ indicates that this is Bruelheide's binomial approach. The corresponding $u$ value will be denoted as $u_{b i n B}$. If successes occur more frequently than expected, $u_{b i n B}$ will be positive. This indicates that relevés containing the species are found to belong to the vegetation unit more often than would be expected by chance. This represents a degree of fidelity of the species to the vegetation unit.

There is an alternative binomial approach. We may take the set of $N_{p}$ relevés belonging to the vegetation unit as our trials and let success be the event that a relevé contains the species in question. The probability of success estimated from the data is $P=n / N$. Here, we assume that $N_{p}$ and the frequency of the species within the data set $(n / N)$ are fixed. The standard deviation is

$$
\begin{equation*}
\sigma_{b i n A}=\sqrt{N_{p} \cdot(n / N) \cdot(1-n / N)} \tag{5}
\end{equation*}
$$

where the subscript $\operatorname{bin} A$ denotes that this is the alternative binomial approach. The corresponding $u$ value will be denoted as $u_{\text {binA }}$. If $u_{\text {binA }}$ is positive, the species occurs in the vegetation unit more often than would be expected by chance. This represents a degree of fidelity of the vegetation unit to the species. Note that in either binomial case, as with the hypergeometric case, the
expected value $\mu=K \cdot P=n \cdot N_{p} / N$.
For completeness, we present here the formula for the standard deviation of a hypergeometric random variable as it is presented in Sokal \& Rohlf (1995: 94):

$$
\begin{equation*}
\sigma_{h y p}=\sqrt{K \cdot P \cdot(1-P) \cdot(N-K) /(N-1)} \tag{6}
\end{equation*}
$$

Regardless of whether the definitions of trial and success are chosen according to Bruelheide's approach or the alternative approach, we get the value of $\sigma_{\text {hyp }}$ given in Eq. 2. Because $u_{\text {hyp }}$ is a measure of joint fidelity, the roles of the vegetation unit and the species can be reversed without affecting the $u_{h y p}$ value.

The binomial $u$ values are related to $u_{h y p}$ by the following equations:
$u_{\text {binB }}=u_{\text {hyp }} \cdot \sqrt{(N-n) /(N-1)}$
$u_{\text {binA }}=u_{\text {hyp }} \cdot \sqrt{\left(N-N_{p}\right) /(N-1)}$

## Phi coefficient

The value $u_{h y p}$ is a measure of statistical significance, and thus it depends upon $N$, the number of relevés in the data set. More relevés give the results greater statistical significance and $u_{\text {hyp }}$ is larger. As an example, consider the case of perfect joint fidelity: the species occurs exclusively within the vegetation unit and every releve from the vegetation unit contains the species. In this case $n_{p}=n=N_{p}$. Calculation shows that $u_{\text {hyp }}=\sqrt{N-1}$, regardless of the value of $n_{p}$. Perfect joint fidelity is given greater value if the data set is larger.

To compare $u_{\text {hyp }}$ values from data sets of different sizes, we can normalize by dividing by the maximum value $u_{\text {hyp }}$ can achieve within the database:

$$
\begin{equation*}
\Phi=\frac{u_{\text {hyp }}}{\sqrt{N-1}}=\frac{N \cdot n_{p}-n \cdot N_{p}}{\sqrt{n \cdot N_{p} \cdot(N-n) \cdot\left(N-N_{p}\right)}} \tag{9}
\end{equation*}
$$

This quantity is the phi coefficient of association (Sokal \& Rohlf 1995: 741, 743). It is independent of the size of the data set. It takes values from -1 to +1 . Positive values indicate that the species and the vegetation unit co-occur more frequently than would be expected by chance. Larger values indicate a greater degree of joint fidelity. The value 1 indicates that the species and the vegetation unit are completely faithful to each other.

## chi-square statistic

The phi coefficient is closely related to the chisquare statistic (Sokal \& Rohlf 1995: 697, 736). Computed for our $2 \times 2$ contingency tables above, the chisquare statistic can be reduced to the following form:
$X^{2}=\sum \frac{\left(f(o)_{i}-f(e)_{i}\right)^{2}}{f(e)_{i}}=\frac{N \cdot\left(N \cdot n_{p}-n \cdot N_{p}\right)^{2}}{n \cdot N_{p} \cdot(N-n) \cdot\left(N-N_{p}\right)}$
Comparing $X^{2}$ with $u_{\text {hyp }}$,
$X^{2}=u_{h y p}^{2} \cdot N /(N-1)$
we see that the two statistics are essentially the same. An advantage of $u_{\text {hyp }}$ over $X^{2}$ is that, whereas $X^{2}$ will take on positive values if the species and vegetation unit cooccur more often or less often than is expected, $u_{\text {hyp }}$ distinguishes between positive and negative fidelity by taking positive values in the former case and negative values in the latter.

## G statistic

The $G$ statistic is an alternative to the chi-square statistic (Sokal \& Rohlf 1995; Botta-Dukát \& Borhidi 1999). Here, we express it in a computational form:
$G=2 \sum f(o)_{i} \cdot \ln \left[f(o)_{i} / f(e)_{i}\right]$

## Estimating probabilities and continuity corrections

If the species and the vegetation unit occur independently, both $X^{2}$ and $G$ statistics have, approximately, a $X^{2}$ distribution with one degree of freedom. The hypergeometric distribution of $u_{\text {hyp }}$ can be approximated by a standard normal distribution. If the statistics are sufficiently large, we can reject the null hypothesis of independence. For example, the probability that a normally distributed random variable will take on a value more than 1.96 standard deviations from the mean is less than $5 \%$. Because $u_{\text {hyp }}$ measures how many standard deviations $n_{p}$ is distant from what would be expected if the species and the vegetation unit were independent (Eq. 1), we can say that values of $\left|u_{\text {hyp }}\right|>1.96$ are statistically significant at $P<0.05$.

The binomial $u$ values will also have, approximately, a normal distribution under the appropriate null hypothesis. For $u_{b i n B}$, the null hypothesis is that among relevés containing the species, relevés belonging to the vegetation unit occur with probability $P=N_{p} / N$. For $u_{\text {binA }}$, the null hypothesis is that among relevés belonging to the vegetation unit, relevés containing the species occur with probability $P=n / N$.

These statistics are often used with a continuity correction to get more accurate estimates of the probability ( $P$-value) that the observed results could have occurred by chance under the given null hypothesis. In the case of $u$ values, Bruelheide $(1995,2000)$ used the following correction:

If $n_{p}-n \cdot N_{p} / N>0.5$, then $u$ adj $=u-0.5 / \sigma$.
If $n_{p}-n \cdot N_{p} / N<-0.5$, then $u$ adj $=u+0.5 / \sigma$.
If $\left|n_{p}^{p}-n \cdot N_{p}^{p} / N\right| \leq 0.5$, then $u$ adj $=0$.
Note that this correction can be substantial when $\sigma$ is small.
$X^{2}$ can be adjusted by the Yates correction (Sokal \& Rohlf 1995: 737):

$$
\begin{equation*}
X_{a d j}^{2}=\frac{N \cdot\left(\left|N \cdot n_{p}-n \cdot N_{p}\right|-(N / 2)\right)^{2}}{n \cdot N_{p} \cdot(N-n) \cdot\left(N-N_{p}\right)} \tag{13}
\end{equation*}
$$

For the $G$ statistic, the Williams correction can be used (Sokal \& Rohlf 1995: 731):

$$
G_{a d j}=\frac{G}{1+\frac{1}{6 N}\left(\frac{N}{n}+\frac{N}{N-n}-1\right)\left(\frac{N}{N_{p}}+\frac{N}{N-N_{p}}-1\right)}(14)
$$

## Fisher's exact test

With modern computing machinery, it is not necessary to resort to approximations and continuity corrections. Fisher's exact test for a right-tail hypothesis precisely calculates the probability of obtaining $f(o)_{l} \geq n_{p}$ (Sokal \& Rohlf 1995: 730, 733). The calculation is based on the hypergeometric distribution. The smaller the calculated probability, the higher the fidelity.

$$
\begin{equation*}
P\left(f(o)_{l} \geq n_{p}\right)=\sum \frac{n!\cdot N_{p}!\cdot(N-n)!\cdot\left(N-N_{p}\right)!}{i!\cdot N!\cdot(n-i)!\cdot\left(N_{p}-i\right)!\cdot\left(N-N_{p}-n+i\right)!} \tag{15}
\end{equation*}
$$

where the sum is taken over all $i \geq n_{p}$. In larger phytosociological data sets, Fisher's exact test may yield very small probability values, including those smaller than $10^{-100}$, which are difficult to cope with in practical work. For this reason, $-\log _{10} \mathrm{P}\left(f(o)_{1} \geq n_{p}\right)$ is a more practical quantity to use for a measure.

## Dufrêne-Legendre Indicator Value Index

Dufrêne \& Legendre (1997; see also Legendre \& Legendre 1998: 369) proposed the Indicator Value Index (IndVal) which is also suitable for determination of fidelity. Unlike the above measures, it is not derived from the comparison of observed and expected frequen-
cies. A potential advantage of this index is that it considers species abundances; the disadvantage is that it ranges from 0 to 1 , not distinguishing negative fidelity. Because it is becoming increasingly popular in ecology, we include it here in its categorical form for comparison. In Dufrêne \& Legendre's (1997) use of this index, fidelity of a species to a vegetation unit is dependent on the delimitation of other vegetation units in the rest of the data set. In our case, we determine species fidelity by comparing the vegetation unit with the rest of the data set as a whole, disregarding any information about the partitions of the data set outside the vegetation unit in question. In this case, using our symbols, the Indicator Value Index for categorical data is expressed as

IndVal $=\frac{n_{p}\left(N-N_{p}\right)}{n \cdot N_{p}-2 n_{p} \cdot N_{p}+n_{p} \cdot N} \cdot \frac{n_{p}}{N_{p}}$

## Material and Methods

To test the performance of statistical fidelity measures, we used a data set of 502 relevés of rock-outcrop dry grasslands (alliances Festucion pallentis and Diantho lumnitzeri-Seslerion) of the Czech Republic, classified into 8 vegetation units (Chytrý et al. unpubl.), called AH in this paper. Classification was performed by the Cocktail method (Bruelheide 1995, 2000). The number of relevés assigned to particular vegetation units has a wide range, from 11 to 204, which is a typical situation for classified relevé data sets. In some analyses, we added 15487 relevés of different types of herbaceous vegetation from the Czech National Phytosociological Database (Chytrý 1997) in order to investigate properties of different fidelity measures in data sets of larger size and different structure. These additional relevés include dry grasslands on deeper soils (not on rock outcrops), meadows, pastures, mountain grasslands, ruderal vegetation, wet grasslands and some other vegetation types. Only vascular plant records were considered because cryptogams were not sampled in some relevés.

Fidelity measures discussed in the previous section were calculated for all species and each of the 8 vegetation units by the computer program JUICE (Tichý 2001), which is freely available at the web site http:// www.sci.muni.cz/botany/juice.htm. The statistics $X^{2}$, $G$ and all three types of $u$ were each calculated both with and without continuity correction. Where continuity correction was applied, the measure is denoted as adjusted (adj). For each vegetation unit, species were ranked by decreasing fidelity, and species ranks produced by different measures were compared. Cluster
analysis (UPGMA, Chord Distance) of the fidelity measures based on standardized species ranks was calculated using the PC-ORD 4 package (McCune \& Mefford 1999). The species ranks were standardized by dividing each rank value by the sum of all rank values for that species. Calculations were performed only with those species which were ranked among the 20 most faithful by at least one fidelity measure.

To demonstrate how fidelity calculations can be used for species ordering in phytosociological tables and to investigate the effect of data set structure on fidelity, we prepared two synoptic tables using JUICE. The first table (Table 5) includes only 502 relevés of the rock outcrop-dry grasslands, each belonging to one of the 8 vegetation units. The second table (Table 6) includes the same relevés and 15487 additional relevés. For defining diagnostic species in these tables, the phi coefficient was used. Threshold $\Phi$ values for species to be diagnostic were arbitrarily set so as to yield 50 diagnostic species in each table. The threshold values were 0.417 for Table 5 and 0.154 for Table 6 . For the latter data set, a lower threshold was necessary because the fidelity of several species decreased due to their occurrence in the additional relevés. The tables were shortened by deleting species without diagnostic capacity.


Fig. 1. Dendrogram based on standardized species ranks resulting from different fidelity measures (see Tables 1 and 2), determined for the vegetation unit A (204 relevés) in the data set of 502 relevés (a) and in the data set of 15898 relevés (b). The distance measure is Relative Euclidean (Chord Distance), the group linkage method is Group Average (UPGMA). The letter $a$ denotes measures applied with the continuity correction.

Table 1. Ranks of species by decreasing fidelity with respect to vegetation unit A ( 204 relevés), calculated by different fidelity measures in the data set of 502 relevés. Only species ranked among the 20 most faithful according to at least one measure are included and species not ranked among 80 most faithful according to at least one measure are excluded. $n=$ number of species occurrences in the entire data set, $n_{p}=$ number of species occurrences in the vegetation unit.

|  | $u_{\text {hyp }}$ | $\begin{gathered} u_{\text {hyp }} \\ \text { adj } \end{gathered}$ | $u_{\text {bin } B}$ | $\begin{gathered} u_{b i n B} \\ a d j \end{gathered}$ | $u_{\text {binA }}$ | $\begin{gathered} u_{b i n A} \\ a d j \end{gathered}$ | $\Phi$ | $X^{2}$ | $\begin{aligned} & X^{2} \\ & a d j \end{aligned}$ | $G$ | $\begin{gathered} G \\ a d j \\ \hline \end{gathered}$ | Fisher | Ind <br> Val | $n$ | $n_{p}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Asplenium septentrionale | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 2 | 190 | 154 |
| Aurinia saxatilis | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 1 | 205 | 160 |
| Hieracium pallidum | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 8 | 101 | 81 |
| Sedum reflexum | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 6 | 144 | 103 |
| Thymus pulegioides | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 7 | 126 | 89 |
| Artemisia campestris | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 4 | 202 | 122 |
| Hieracium cymosum | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 23 | 63 | 44 |
| Pulsatilla pratensis | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 10 | 10 | 10 | 19 | 76 | 50 |
| Festuca pallens | 8 | 8 | 29 | 26 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 3 | 334 | 161 |
| Veronica dillenii | 10 | 10 | 8 | 8 | 10 | 10 | 10 | 10 | 10 | 9 | 9 | 9 | 38 | 27 | 23 |
| Hieracium umbellatum | 11 | 11 | 10 | 10 | 11 | 11 | 11 | 11 | 11 | 11 | 11 | 11 | 33 | 46 | 33 |
| Galium glaucum | 12 | 12 | 21 | 19 | 12 | 12 | 12 | 12 | 12 | 12 | 12 | 12 | 9 | 217 | 112 |
| Campanula rotundifolia agg. | 13 | 13 | 11 | 11 | 13 | 13 | 13 | 13 | 13 | 13 | 13 | 13 | 20 | 82 | 51 |
| Verbascum lychnitis | 15 | 14 | 15 | 15 | 15 | 14 | 15 | 15 | 14 | 15 | 15 | 14 | 17 | 98 | 57 |
| Hieracium sabaudum | 14 | 15 | 12 | 12 | 14 | 15 | 14 | 14 | 15 | 14 | 14 | 15 | 29 | 59 | 38 |
| Rumex acetosella agg. | 17 | 16 | 16 | 16 | 17 | 16 | 17 | 17 | 16 | 17 | 17 | 16 | 22 | 84 | 50 |
| Stipa joannis | 16 | 17 | 13 | 13 | 16 | 17 | 16 | 16 | 17 | 16 | 16 | 17 | 36 | 46 | 31 |
| Lychnis viscaria | 18 | 18 | 14 | 14 | 18 | 18 | 18 | 18 | 18 | 18 | 18 | 18 | 45 | 23 | 18 |
| Potentilla neumanniana | 19 | 19 | 18 | 17 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 35 | 51 | 33 |
| Anthericum liliago | 21 | 20 | 20 | 20 | 21 | 20 | 21 | 21 | 20 | 21 | 21 | 20 | 28 | 70 | 42 |
| Hieracium pilosella | 22 | 21 | 24 | 24 | 22 | 21 | 22 | 22 | 21 | 22 | 22 | 22 | 18 | 105 | 58 |
| Hieracium bauhinii | 20 | 22 | 17 | 18 | 20 | 22 | 20 | 20 | 22 | 20 | 20 | 21 | 55 | 17 | 14 |
| Artemisia absinthium | 25 | 25 | 19 | 22 | 25 | 25 | 25 | 25 | 25 | 24 | 24 | 25 | 52 | 21 | 16 |
| Dianthus carthusianorum agg. | 28 | 28 | 35 | 34 | 28 | 28 | 28 | 28 | 28 | 29 | 29 | 28 | 11 | 179 | 89 |
| Euphorbia cyparissias | 27 | 27 | 49 | 42 | 27 | 27 | 27 | 27 | 27 | 28 | 28 | 27 | 5 | 324 | 148 |
| Sedum maximum | 46 | 44 | 48 | 46 | 46 | 44 | 46 | 46 | 44 | 46 | 44 | 44 | 15 | 151 | 72 |
| Allium montanum | 79 | 74 | 80 | 77 | 79 | 74 | 79 | 79 | 74 | 79 | 78 | 77 | 10 | 271 | 113 |

## Results and Discussion

## Empirical comparison of fidelity measures

Species ranks yielded by different fidelity measures for a single vegetation unit are compared in Fig. 1 and Tables 1 and 2. Two cases are shown: a vegetation unit which is (a) large and (b) small relative to the size of the data set. Both cases are represented by vegetation unit A ( 204 relevés), treated in the data sets of 502 and 15989 relevés, respectively.

Due to the relationships among them, $u_{\text {hyp }}, u_{\text {binA }}, \Phi$, and $X^{2}$ yield exactly the same species ranks in both cases. If $u_{\text {hyp }}, u_{\text {binA }}$, and $X^{2}$ are corrected for continuity, the ranks for adjusted measures are again identical, but differ from the ranks yielded with unadjusted measures. This is due to the tendency of adjusted measures to give slightly lower fidelity values to rare species. Species ranks produced by $u_{\text {binB }}$ and $G$ remarkably differ from those produced by the above measures and have also a low similarity to one another. In $u_{b i n B}$, this difference is because it tends to underweight fidelity of common species (e.g. Euphorbia cyparissias and Dianthus carthusianorum agg. in Tables 1 and 2). This is in accordance with Eq. 7, which shows that $u_{\text {binB }}$ deviates from $u_{\text {hyp }}$ (and related measures) unless the number of species occurrences in the data set ( $n$ ) is small
relative to the total number of relevés $(N)$. By contrast, the $G$ statistic tends to underweight fidelity of rare species, particularly if the vegetation unit is small relative to the data set size (e.g. Hieracium cymosum, H. sabaudum, and Asplenium trichomanes in Table 2). The same, but a much stronger tendency was found in the categorical form of the Indicator Value Index which gives the most deviating results compared with the other measures (note that it yields high values for common species such as Euphorbia cyparissias, Artemisia campestris, and Dianthus carthusianorum agg. in Tables 1 and 2).

## Joint fidelity measures

Botta-Dukát \& Borhidi (1999) emphasized the distinction between the joint fidelity and two asymmetric fidelity measures. We believe that the traditional understanding of phytosociological fidelity is close to the joint fidelity and, to some extent, also to the fidelity of a species to a vegetation unit. First, we discuss properties of the joint fidelity measures, i.e. $u_{\text {hyp }}, \Phi, X^{2}, G$ and Fisher's exact test. Table 3 compares these measures in six test cases. Cases 1,3 and 5 represent perfect joint fidelity, i.e. any relevé belonging to the vegetation unit contains the species and any relevé containing the species belongs to the vegetation unit. In Cases 2, 4 and 6 fidelity is still high

Table 2. Ranks of species by decreasing fidelity with respect to vegetation unit A ( 204 relevés), calculated by different fidelity measures in the data set of 15898 relevés. Only species ranked among the 20 most faithful according to at least one measure are included. $n=$ number of species occurrences in the entire data set, $n_{p}=$ number of species occurrences in the vegetation unit.

|  | $u_{\text {hyp }}$ | $\begin{gathered} u_{\text {hyp }} \\ \text { adj } \end{gathered}$ | $u_{\text {binB }}$ | $\begin{gathered} u_{b i n B} \\ a d j \end{gathered}$ | $u_{\text {binA }}$ | $\begin{gathered} u_{b i n A} \\ \text { adj } \end{gathered}$ | $\Phi$ | $X^{2}$ | $\begin{gathered} X^{2} \\ a d j \end{gathered}$ | $G$ | $\begin{gathered} G \\ a d j \end{gathered}$ | Fisher | Ind <br> Val | $n$ | $n_{p}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Asplenium septentrionale | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 2 | 2 | 2 | 2 | 279 | 154 |
| Aurinia saxatilis | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 1 | 1 | 1 | 1 | 307 | 160 |
| Hieracium pallidum | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 4 | 5 | 4 | 12 | 133 | 81 |
| $F e s t u c a ~ p a l l e n s ~$ | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 3 | 3 | 3 | 3 | 891 | 161 |
| Sedum reflexum | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 5 | 4 | 5 | 8 | 353 | 103 |
| Allium montanum | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 6 | 578 | 113 |
| Galium glaucum | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 7 | 7 | 7 | 7 | 841 | 112 |
| Sedum album | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 10 | 10 | 10 | 10 | 735 | 95 |
| Artemisia campestris | 10 | 10 | 12 | 12 | 10 | 10 | 10 | 10 | 10 | 8 | 8 | 8 | 5 | 1265 | 122 |
| Hieracium cymosum | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 14 | 14 | 14 | 25 | 137 | 44 |
| Seseli osseum | 11 | 11 | 11 | 11 | 11 | 11 | 11 | 11 | 11 | 9 | 9 | 9 | 9 | 953 | 105 |
| Jovibarba sobolifera | 12 | 12 | 10 | 10 | 12 | 12 | 12 | 12 | 12 | 13 | 13 | 13 | 19 | 306 | 57 |
| Sedum maximum | 13 | 13 | 13 | 13 | 13 | 13 | 13 | 13 | 13 | 12 | 12 | 12 | 16 | 558 | 72 |
| Melica transsilvanica | 14 | 14 | 14 | 14 | 14 | 14 | 14 | 14 | 14 | 15 | 16 | 15 | 22 | 411 | 56 |
| Vincetoxicum hirundinaria | 16 | 16 | 16 | 16 | 16 | 16 | 16 | 16 | 16 | 16 | 15 | 16 | 18 | 568 | 63 |
| Euphorbia cyparissias | 20 | 20 | 26 | 26 | 20 | 20 | 20 | 20 | 20 | 11 | 11 | 11 | 4 | 3376 | 148 |
| Pulsatilla pratensis | 17 | 17 | 18 | 17 | 17 | 17 | 17 | 17 | 17 | 20 | 20 | 20 | 24 | 430 | 50 |
| Asplenium trichomanes | 15 | 15 | 15 | 15 | 15 | 15 | 15 | 15 | 15 | 23 | 23 | 23 | 35 | 156 | 133 |
| Verbascum lychnitis | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 19 | 23 | 597 | 57 |
| Hieracium sabaudum | 18 | 18 | 17 | 18 | 18 | 18 | 18 | 18 | 18 | 24 | 25 | 24 | 32 | 260 | 38 |
| Anthericum liliago | 21 | 21 | 20 | 20 | 21 | 21 | 21 | 21 | 21 | 26 | 26 | 26 | 29 | 352 | 42 |
| Thymus pulegioides | 26 | 25 | 27 | 27 | 26 | 25 | 26 | 26 | 25 | 18 | 18 | 18 | 13 | 1526 | 89 |
| Potentilla arenaria | 28 | 27 | 28 | 28 | 28 | 27 | 28 | 28 | 27 | 17 | 17 | 17 | 11 | 1980 | 101 |
| Dianthus carthusianorum agg. | 33 | 32 | 36 | 34 | 33 | 32 | 33 | 33 | 32 | 21 | 21 | 21 | 14 | 1803 | 89 |
| Asperula cynanchica | 35 | 35 | 37 | 35 | 35 | 35 | 35 | 35 | 35 | 22 | 22 | 22 | 15 | 1799 | 88 |
| Echium vulgare | 34 | 33 | 34 | 33 | 34 | 33 | 34 | 34 | 33 | 28 | 28 | 28 | 20 | 1022 | 64 |

but not perfect, with the target species being absent in some relevés of the vegetation unit and present in some relevés outside the vegetation unit.

By increasing the number of relevés in the vegetation unit $\left(N_{p}\right)$ from 20 to 200 while maintaining the perfect joint fidelity (Cases 1 and 3 ) the $P$-value calculated by Fisher's exact test decreases from $10^{-42}$ to $10^{-216}$. Case 4 does not represent perfect joint fidelity, but because the vegetation unit is larger, the results are statistically less probable than the perfect fidelity of Case 1 . That is, the probability of the results occurring by chance is lower in Case 4 than in Case 1. By increasing the size of the entire data set (Cases 5 and 6), the results become even less probable. Intuitively, however, Cases 3 and 5, or Cases 4 and 6, respectively, can be treated as the same. Therefore, probability may not be considered as the most appropriate measure of fidelity in some cases.

The $G$ statistic behaves in the same way as Fisher's exact test. The measures $u_{\text {hyp }}$ and $X^{2}$ are not as dependent on the size of the vegetation unit as Fisher's exact test or the $G$ statistic. However, their values increase with the increasing size of the data set. Therefore they produce roughly comparable results for differently sized vegetation units but not for differently sized data sets.

The phi coefficient is independent of the size of the data set. It is equal to 1 for all cases of perfect joint fidelity and yields roughly comparable values for Cases 2,4 and 6 , which seems more in accordance with intui-
tion. It should be noted, however, that unlike $u_{\text {hyp }}, X^{2}, G$ or Fisher's exact test, the phi coefficient contains no information about statistical significance. For small data sets or small vegetation units, use of the phi coefficient to measure fidelity may lead to invalid conclusions. By contrast, in large data sets the fidelity values for the most faithful species are usually far beyond the conventional significance levels. In the latter case, the phi coefficient is a particularly appropriate fidelity measure as it yields comparable values among differently sized vegetation units and among differently sized data sets.

Table 3. Comparison of the joint fidelity measures in six test cases, with Cases 1, 3 and 5 being the perfect joint fidelity of the species and the vegetation unit to each other. $N=$ number of relevés in the data set; $N_{p}=$ number of relevés in the vegetation unit; $n=$ number of species occurrences in the data set; $n_{p}=$ number of species occurrences in the vegetation unit.

|  | Case 1 | Case 2 | Case 3 | Case 4 | Case 5 | Case 6 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| $N$ | 1000 | 1000 | 1000 | 1000 | 10000 | 10000 |
| $N_{p}$ | 20 | 20 | 200 | 200 | 2000 | 2000 |
| $n$ | 20 | 20 | 200 | 200 | 2000 | 2000 |
| $n_{p}$ | 20 | 15 | 200 | 150 | 2000 | 1500 |
|  |  |  |  |  |  |  |
| $u_{\text {hyp }}$ | 31.61 | 23.54 | 31.61 | 21.73 | 100.00 | 68.75 |
| $\Phi$ | 1 | 0.74 | 1 | 0.69 | 1 | 0.69 |
| $X^{2}$ | 1000 | 555 | 1000 | 473 | 10000 | 4727 |
| $G$ | 196 | 111 | 1000 | 402 | 10000 | 4018 |
| Fisher | $10^{-42}$ | $10^{-24}$ | $10^{-216}$ | $10^{-88}$ | $<10^{-1000}$ | $10^{-874}$ |

It is important to note that, within a given database, the phi coefficient and $u_{\text {hyp }}$ will give the same ranks of fidelities, although the values will not be the same (Eq. 9). The $X^{2}$ statistic will also yield the same results (Eq. 11), assuming we ignore the pairs which are negatively faithful to each other.

## Asymmetric fidelity measures

Botta-Dukát \& Borhidi (1999) derived three variants of the $G$ statistic to measure joint fidelity, fidelity of the species to the vegetation unit and fidelity of the vegetation unit to the species. These three forms of fidelity can be also measured, in turn, by $u_{\text {hyp }}, u_{\text {binB }}$ and $u_{\text {binA }}$.

Table 4 shows three test cases, representing the three forms of fidelity. In Case 1 the species and the vegetation unit are perfectly jointly faithful to each other. In Case 2 the species is faithful to the vegetation unit, but the vegetation unit is not faithful to the species. Case 3 is the opposite of Case 2. All measures have high values in Case 1. Cases 2 and 3 cannot be distinguished by $u_{\text {hyp }}$ because it is a joint fidelity measure. By contrast, $u_{b i n B}$ yields higher values in Case 2 than in Case 3, which demonstrates its suitability for measuring the fidelity of a species to a vegetation unit. The opposite pattern is shown for $u_{\text {binA }}$, which indicates that it gives a higher weight to the fidelity of a vegetation unit to a species.

In Tables 1 and 2 (see also Fig. 1), $u_{\text {binA }}$ yields the same species ranks for one vegetation unit as $u_{\text {hyp }}, X^{2}$, and $\Phi$. This is explained by Eq. 8 which shows that $u_{\text {binA }}$ is related to these measures through variables $N$ and $N_{p}$ which are the same for all species, but not through $n$ or $n_{p}$ which vary among species. If we theoretically ranked vegetation units by their fidelity to a single species, identical ranks would be yielded for $u_{\text {hyp }} X^{2}, \Phi$, and $u_{\text {binB }}$, whereas $u_{\text {binA }}$ would deviate (Eq. 7).

In conclusion, $u_{\text {binB }}$ can also be used as an appropriate fidelity measure, provided there is an intention to give a higher value to the phenomenon of species-to-vegetation unit fidelity, i. e. to highlight the species which occur in few relevés of a vegetation unit due to their overall rarity, but are hardly found outside this unit. By contrast $u_{\text {binA }}$ does not have a practical value for measuring fidelity. If strong downweighting of fidelity of rare species is required, which is perhaps rarely the aim in phytosociology, the categorical form of the Dufrêne-Legendre Indicator Value Index can be used. This index, however, may be advantageous when working with abundance/cover values.

## Fidelity-based sorting of species in phytosociological tables

Phytosociological tables are a common tool for visualization of vegetation data (Westhoff \& van der

Maarel 1973; Dierschke 1994). Species sorting in these tables is usually done by intuitive comparison of differences in species frequencies among vegetation units. Following the ideas pioneered in recent studies based on $u$ values (Bruelheide 1995; Bruelheide \& Jandt 1995, 1997; Jandt 1999; Pflume 1999; Bruelheide \& Chytrý 2000; Täuber 2000), we suggest an automatic procedure for fidelity-based species sorting in phytosociological tables, which is available in the JUICE program (Tichý 2001). In the test examples of Tables 5 and 6 , we demonstrate this new method on synoptic tables sorted by the phi coefficient. It can be also applied in combination with any other fidelity measure and to any table of individual relevés which has been previously partitioned by some classification method.

In phytosociological tables, diagnostic species for particular vegetation units are usually clustered into diagonally arranged blocks. Diagnostic species to be included into these blocks can be defined as those exceeding some arbitrary threshold value of fidelity. The lower the threshold value is set, the more species will be considered as diagnostic and the larger proportion of the table's species will be assigned to the blocks. In Tables 5 and 6, the aim was to have 50 diagnostic species in each and the thresholds were selected accordingly.

As the aim of species blocks is displaying differentiation of vegetation units, it is reasonable to rank species within blocks by decreasing fidelity (i.e. differentiation capacity) rather than by decreasing percentage frequency which is the option commonly used in traditional tables. Tables 5 and 6 are sorted by fidelity. They clearly show that there is no perfect correlation between species fidelity and frequency.

Properties of the fidelity measures determine an important feature of Tables 5 and 6 which has never been considered in the traditional tables prepared intuitively by experts. Some species which are relatively common in the data set do not have diagnostic value for vegetation

Table 4. Comparison of the performance of $u_{\text {hyp }}, u_{\text {binB }}$, and $u_{\text {binA }}$ in three test cases, representing different forms of fidelity. $N, N_{p}, n$ and $n_{p}$ as in Table 3.

|  | Case 1 | Case 2 | Case 3 |
| :--- | ---: | ---: | ---: |
| Joint fidelity <br> Fidelity of the species <br> to the vegetation unit <br> Fidelity of the vegetation <br> unit to the species <br> $N$ | High | High | High |
| $N_{p}$ | High | Lorate | Moderate |
| Low |  |  |  |

Table 5. Synoptic table of 502 relevés of the Czech rock-outcrop dry grasslands, based on fidelity comparison without additional relevés. Diagnostic species (values grey-shaded) are those with $\Phi \geq 0.417$; they are ranked by decreasing value of $\Phi$. Dots in part (a) indicate species absence, dashes in part (b) of the table indicate negative fidelity. Asterisks before species names indicate species which are also diagnostic in Table 6.

| Vegetation unit | (a) percentage frequency |  |  |  |  |  |  |  | (b) phi coefficient $\times 1000$ |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | A | B | C | D | E | F | G | H | A | B | C | D | E | F | G | H |
| Nr. of relevés | 204 | 78 | 25 | 30 | 66 | 73 | 11 | 15 | 204 | 78 | 25 | 30 | 66 | 73 | 11 | 15 |
| Nr. of diagnostic species | 2 | 0 | 13 | 14 | 2 | 2 | 8 | 13 | 2 | 0 | 13 | 14 | 2 | 2 | 8 | 13 |
| Diagnostic species of the vegetation unit $A$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Asplenium septentrionale | 75 | 23 | . | . | 3 | 22 | . | . | 642 | - | - | - | - | - | - | - |
| *Aurinia saxatilis | 78 | 36 | . | . | 9 | 15 | . | . | 633 | - | - | - | - | - | - | - |
| Diagnostic species of the vegetation unit C |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Achillea setacea | . | 3 | 48 | . | . | . | . | . | - | - | 629 | - | - | - | - | - |
| Agrostis stricta | 2 | 1 | 56 | . | . | 1 | . | . | - | - | 609 | - | - | - | - | - |
| *Helichrysum arenarium | 1 | . | 40 | . | . | . | . | . | - | - | 564 | - | - | - | - | - |
| Jasione montana | 5 | . | 56 | . | . | . | . | . | - | - | 550 | - | - | - | - | - |
| Carex supina | . | . | 28 | . | . | . | . | . | - | - | 519 | - | - | - | - | - |
| Rumex acetosella agg. | 25 | 6 | 100 | . | . | . | 36 | . | 172 | - | 511 | - | - | - | 79 | - |
| Erophila verna agg. | 0 | 3 | 36 | . | 2 | . | . | . | - | - | 482 | - | - | - | - | - |
| *Hieracium echioides | 6 | 3 | 56 | . | 5 | . | . | . | - | - | 474 | - | - | - | - | - |
| Avenochloa pratensis | 2 | . | 52 | . | 3 | 5 | 45 | . | - | - | 463 | - | - | - | 260 | - |
| *Gagea bohemica | 0 |  | 24 | . | . |  | . | . | - | - | 441 | - | - | - | - | - |
| *Scleranthus perennis | 13 |  | 64 | . |  | 3 | . | . | 124 | - | 441 | - | - | - | - | - |
| Diagnostic species of the vegetation unit $D$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Teucrium montanum |  | . | . | 97 | 5 | 3 | . | . | - | - | - | 902 | - | - | - | - |
| *Scorzonera austriaca | . | . | . | 73 | 6 | 1 | . | . | - | - | - | 759 | 12 | - | - | - |
| * Poa badensis |  | 4 | . | 80 | 8 | 4 | . | . | - | - | - | 723 | - | - | - | - |
| * Campanula sibirica | . | . | . | 63 | 5 | 3 | . | . | - | - | - | 692 | - | - | - | - |
| *Fumana procumbens | . | . | $\cdot$ | 53 | 2 | 1 | . | . | - | - | - | 675 | - | - | - | - |
| *Minuartia setacea | 1 | 4 | 4 | 100 | 18 | 23 | . | . | - | - | - | 648 | 58 | 124 | - | - |
| *Melica ciliata | 1 | 5 | . | 77 | 8 | 15 | . | . | - | - | - | 597 | - | 88 | - | - |
| * Dorycnium germanicum | . | . | 4 | 70 | 8 | 7 | 55 | . | - | - | - | 595 | - | - | 266 | - |
| *Allium flavum | 3 | 3 | 8 | 77 | 14 | 11 | 45 | . | - | - | - | 525 | 31 | - | 163 | - |
| Medicago falcata |  | 4 | 4 | 47 | 6 |  | . | 7 | - | - | - | 507 | 28 | - | - | 18 |
| *Alyssum montanum | 7 | 14 | 8 | 87 | 24 | 15 | 18 | . | - | - | - | 480 | 83 | - | - | - |
| Stipa capillata | 6 | 8 | . | 60 | 8 | 1 | . | . | - | - | - | 463 | - | - | - | - |
| Astragalus austriacus | . | . | . | 33 | 3 | 4 | . | . | - | - | - | 449 | - | 27 | - | - |
| Inula ensifolia | . | . | . | 33 | 6 | 4 | . | . | - | - | - | 417 | 58 | 16 | - | - |
| Diagnostic species of the vegetation unit E |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Helianthemum canum |  | 3 | - | 3 | 32 | 5 | . | 7 | - | - | - | - | 434 | - | - | - |
| Diagnostic species of the vegetation unit F |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Saxifraga paniculata |  |  | . | . | 18 | 41 | . | . | - | - | - | - | 138 | 488 | - | - |
| Diagnostic species of the vegetation unit G |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Asplenium cuneifolium | . | . | - | . | . | . | 100 | . | - | - | - | - | - | - | 1000 | - |
| *Thlaspi montanum |  | . | . | . | 3 | . | 73 | . | - | - | - | - | 29 | - | 758 | - |
| Rumex acetosa |  | . | . |  | 2 | 1 | 55 | . | - | - | - | - | - | - | 633 | - |
| * Biscutella laevigata | 1 | 3 | 12 | 3 | 14 | 19 | 100 | . | - | - | 28 | - | 70 | 156 | 489 | - |
| Festuca ovina | 7 | 5 | 12 | . | 5 | 5 | 82 | . | - | - | 41 | - | - | - | 426 | - |
| Diagnostic species of the vegetation unit H |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| * Cirsium acaule |  | . | . | . | . | . | . | 87 | - | - | - | - | - | - | - | 929 |
| Ononis spinosa | . | . | . | . | 3 | 1 | . | 80 | - | - | - | - | - | - | - | 794 |
| * Coronilla vaginalis | . | . | . | . | . | . | . | 40 | - | - | - | - | - | - | - | 627 |
| Brachypodium pinnatum | 2 | . | . | . | 12 | 7 | . | 93 | - | - | - | - | 92 | - | - | 625 |
| Inula salicina | . | 1 | . | . | . | . | . | 40 | - | - | - | - | - | - | - | 578 |
| Gentianella ciliata | . | . | . | . | . | . | . | 33 | - | - | - | - | - | - | - | 572 |
| Scorzonera hispanica | . | . | . | . | . | . | . | 33 | - | - | - | - | - | - | - | 572 |
| Prunella grandiflora | . | . | . | . | . | . | . | 33 | - | - | - | - | - | - | - | 572 |
| Carex flacca | . | . | . | . | . | . | . | 33 | - | - | - | - | - | - | - | 572 |
| Cirsium pannonicum | . | . | . | . | 2 | 1 | . | 33 | - | - | - | - | - | - | - | 478 |
| Carlina vulgaris agg. | 0 | . | 4 | . | 2 | 5 | . | 47 | - | - | 17 | - | - | 67 | - | 468 |
| Bromus erectus | . | . | . | . | . | . | . | 20 | - | - | - | - | - | - | - | 442 |
| Centaurea jacea | . | . | . | . | . |  |  | 20 | - | - | - | - | - | - | - | 442 |
| Common diagnostic species of two vegetation units |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| * Armeria elongata | 1 | . | 52 | . | . |  | 91 | . | - | - | 495 | - | - | - | 591 | - |
| *Genista pilosa | 0 | . | 60 |  | 2 | 14 | 82 | . | - | - | 469 | - | - | 104 | 433 | - |
| *Sesleria varia | 2 | 1 | . | 40 | 98 | 100 | 100 | 100 | - | - | - | 21 | 506 | 549 | 199 | 234 |

Table 6. Synoptic table of 502 relevés of Czech rock-outcrop dry grasslands, based on fidelity comparison with 15487 additional relevés of various grassland types of the Czech Republic. Diagnostic species (values grey-shaded) are those with $\Phi \geq 0.154$; they are ranked by decreasing value of $\Phi$. Dots in part (a) indicate species absence, dashes in part (b) of the table indicate negative fidelity. Asterisks before species names indicate species which are also diagnostic in Table 5.

| Vegetation unit | (a) percentage frequency |  |  |  |  |  |  |  | (b) phi coefficient $\times 1000$ |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | A | B | C | D | E | F | G | H | A | B | C | D | E | F | G | H |
| Nr . of relevés | 204 | 78 | 25 | 30 | 66 | 73 | 11 | 15 | 204 | 78 | 25 | 0 | 66 | 73 | 11 | 15 |
| Nr. of diagnostic species | 16 | 7 | 8 | 16 | 13 | 14 | 7 | 3 | 16 | 7 | 8 | 16 | 13 | 14 | 7 | 3 |
| Diagnostic species of the vegetation unit A |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Asplenium septentrionale | 75 | 23 |  |  | 3 | 22 |  |  | 640 | 114 | - | - | - | 104 | - | - |
| Hieracium pallidum | 40 | 5 | 4 |  | 8 | 14 | . |  | 487 | 33 | 14 | - | 48 | 96 | - | - |
| Sedum reflexum | 50 | 13 | 24 |  | 6 | 29 | . |  | 373 | 51 | 59 | - | 17 | 122 | - | - |
| Hieracium cymosum | 22 | 3 |  |  | 9 | 15 | . | . | 255 | 13 | - | - | 57 | 104 | - | - |
| Artemisia campestris | 60 | 42 | 44 | 57 | 18 | 10 | . | . | 218 | 89 | 53 | 78 | 24 | - | - | - |
| Diagnostic species of the vegetation unit $\mathbf{C}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Hieracium echioides | 6 | 3 | 56 |  | 5 | . | . | . | 47 | - | 195 | - | 19 | - | - | - |
| *Gagea bohemica | 0 | . | 24 |  | . | . | . | . | - | - | 188 | - | - | - | - | - |
| *Helichrysum arenarium | 1 | . | 40 |  | . |  |  |  | - | - | 167 | - | - | - | - | - |
| *Scleranthus perennis | 13 |  | 64 |  | . | 3 | . | . | 85 | - | 166 | - | - | - | - | - |
| *Achillea setacea |  | 3 | 48 |  | . |  | . |  | - | - | 154 | - | - | - | - | - |
| Diagnostic species of the vegetation unit D |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Teucrium montanum | . | . |  | 97 | 5 | 3 |  | . | - | - | - | 701 | 45 | 27 | - | - |
| *Scorzonera austriaca | . | . | . | 73 | 6 | 1 | . | . | - | - | - | 561 | 66 | 13 | - | - |
| *Fumana procumbens | . |  | . | 53 | 2 | 1 | . | . | - | - | - | 532 | 20 | 18 |  | - |
| *Poa badensis |  | 4 |  | 80 | 8 | 4 | . |  | - | 30 | - | 458 | 60 | 32 | - | - |
| *Minuartia setacea | 1 | 4 | 4 | 100 | 18 | 23 |  | . | - | 20 | 12 | 427 | 110 | 150 | - | - |
| *Melica ciliata | 1 | 5 |  | 77 | 8 | 15 | . | . | - | 29 | - | 331 | 43 | 96 | - | - |
| *Campanula sibirica |  |  |  | 63 | 5 | 3 |  |  | - | - | - | 263 | 22 | 11 |  | - |
| *Allium flavum | 3 | 3 | 8 | 77 | 14 | 11 | 45 |  | 14 | - | 19 | 248 | 58 | 47 | 87 | - |
| *Alyssum montanum | 7 | 14 | 8 | 87 | 24 | 15 | 18 | . | 32 | 52 | 14 | 234 | 90 | 55 | 26 | - |
| Diplotaxis muralis | . | . |  | 7 |  |  |  |  | - | - | - | 182 | - | - | - | - |
| *Dorycnium germanicum | . | . | 4 | 70 | 8 | 7 | 55 | . | - | - | - | 165 | 16 | 14 | 77 | - |
| Diagnostic species of the vegetation unit $E$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Helianthemum canum | . | 3 |  | 3 | 32 | 5 |  | 7 | - | 14 | - | 12 | 224 | 36 | - | 20 |
| Stachys recta | 31 | 53 | 4 | 47 | 86 | 18 |  | 7 | 122 | 140 | - | 76 | 222 | 35 | - | - |
| Diagnostic species of the vegetation unit $F$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Cardaminopsis petraea | 0 |  |  |  | 3 | 11 | . |  | 17 | - | - | - | 69 | 269 | - | - |
| Viola collina | 3 |  | . | . | 8 | 16 | . |  | 48 | - | - | - | 77 | 179 | - | - |
| Cardaminopsis arenosa | 10 | 3 | . | . | 18 | 19 |  | . | 137 | 17 | - | - | 141 | 157 | - | - |
| Diagnostic species of the vegetation unit G |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Asplenium cuneifolium | . | . |  |  |  | . | 100 | . | - | - | - | - | - | - | 782 | - |
| *Thlaspi montanum |  | . | . |  | 3 |  | 73 |  | - | - | - | - | 50 | - | 514 | - |
| Stellaria holostea | 1 | . |  | . | 2 | 1 | 36 | . | 19 | - | - | - | 18 | 17 | 206 | - |
| Diagnostic species of the vegetation unit H |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Coronilla vaginalis | . | . | . | . | . |  |  | 40 | - | - | - | - | - | - | - | 225 |
| *Cirsium acaule | . | - | . | . | . |  |  | 87 | - | - | - | - | - | - | - | 168 |
| Common diagnostic species of two vegetation units |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| *Aurinia saxatilis | 78 | 36 | . | . | 9 | 15 |  | . | 634 | 173 | - | - | 34 | 65 | - | - |
| Sedum maximum | 35 | 58 | 4 | . | 17 | 19 | 73 | . | 197 | 207 | - | - | 46 | 58 | 99 | - |
| Melica transsilvanica | 27 | 45 | 4 | 3 | 14 | 14 |  | . | 179 | 187 | - | - | 45 | 48 | - | - |
| Galium glaucum | 55 | 35 | . | 33 | 74 | 22 |  | 20 | 253 | 92 | - | 54 | 199 | 50 | - | 20 |
| Asplenium trichomanes | 16 | 14 | . |  | 17 | 48 | 18 | . | 176 | 93 | - | - | 103 | 324 | 46 | - |
| *Genista pilosa | 0 | . | 60 | . | 2 | 14 | 82 | . | - | - | 206 | - | - | 74 | 187 | - |
| *Armeria elongata | 1 | . | 52 |  | . |  | 91 |  | - | - | 167 | - | - | - | 195 | - |
| Anthericum ramosum | 8 | 10 | 8 | 90 | 56 | 49 |  | 33 | 22 | 20 | - | 183 | 164 | 150 | - | 44 |
| Asplenium ruta-muraria | 14 | 18 |  | 7 | 38 | 47 |  |  | 132 | 109 | - | 22 | 219 | 284 | - | - |
| Cotoneaster integerrimus | 17 | 12 | 4 | . | 38 | 37 | . | . | 140 | 55 | - | - | 186 | 190 | - | - |
| *Saxifraga paniculata |  |  |  | . | 18 | 41 |  |  | - | - | - | - | 173 | 417 | - | - |
| Biscutella laevigata | 1 | 3 | 12 | 3 | 14 | 19 | 100 | . | 13 | 18 | 58 | 15 | 109 | 163 | 338 | - |
| Common diagnostic species of more than two vegetation units |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Allium montanum | 55 | 100 |  | 30 | 56 | 47 | . | . | 315 | 362 | - | 61 | 181 | 156 | - | - |
| Sedum album | 47 | 88 |  | 90 | 64 | 62 | . | . | 228 | 280 | - | 177 | 181 | 184 | - | - |
| Seseli osseum | 51 | 62 | 52 | 100 | 83 | 82 |  | . | 218 | 164 | 77 | 172 | 210 | 218 | - | - |
| Festuca pallens | 79 | 44 | 100 | 100 | 53 | 64 | 18 |  | 363 | 116 | 163 | 178 | 133 | 174 | 14 | - |
| Jovibarba sobolifera | 28 | 33 | 4 | 87 | 38 | 29 |  |  | 216 | 161 | - | 268 | 169 | 133 | - | - |
| Vincetoxicum hirundinaria | 31 | 21 | 4 | 7 | 55 | 68 |  |  | 168 | 64 | - | - | 177 | 238 | - | - |
| *Sesleria varia | 2 | 1 | . | 40 | 98 | 100 | 100 | 100 | - | - | - | 108 | 408 | 436 | 169 | 197 |

units with few relevés, even though they have a high frequency in these vegetation units. For example, Sedum maximum in Table 6 is diagnostic for vegetation units A and B where its frequency is $35 \%$ and $58 \%$, respectively, but is not diagnostic for vegetation unit $G$ where its frequency is $73 \%$ (note that unit A includes 204 relevés, B 78 , and G only 11). A similar pattern is shown by Sesleria varia in Table 5 and Asplenium trichomanes, Seseli osseum and Jovibarba sobolifera in Table 6. This is in accordance with the well-known statistical rule that observed phenomena (in this case species occurrence in a vegetation unit) are more significant if observed in more cases (i.e. in more relevés).

## Determination of diagnostic species in large data sets

During the history of phytosociology, diagnostic species could have hardly been determined from real data because even large tables only included relevés of a few related vegetation types or relevés from restricted areas. Expert judgement was the only way to determine diagnostic species of more general validity before large phytosociological data sets and corresponding computer technology became available. Table 5, compiled from a data set which includes only 502 relevés of the rockoutcrop dry grasslands, represents this traditional approach. In this table diagnostic value is given to several species with broad ecological ranges which are frequently found in many different vegetation types. Many of them have their ecological optima outside the rock outcrops (e.g. Jasione montana, Rumex acetosella, R. acetosa, Festuca ovina, Brachypodium pinnatum, Carex flacca, Bromus erectus and Centaurea jacea). These species can be valuable for differentiating particular vegetation units against the others in the rock-outcrop grassland data set, but not against the units outside the data set.

With currently available phytosociological databases, it is possible to determine diagnostic species in large data sets which include not only relevés of target vegetation types but also relevés of most vegetation types occurring in a wider area. This approach yields diagnostic species of more general validity. Table 6 shows diagnostic species of the same vegetation units as in Table 5, however, these diagnostic species were determined in a data set where several thousands of relevés of different types of herbaceous vegetation were added. It is evident that species with broad ecological ranges are no longer indicated as diagnostic. In addition, ecologically specialized species which occur in many types of rock-outcrop grasslands but are rare in other vegetation types are indicated as diagnostic in Table 6 but not in Table 5. For example, Allium montanum, Sedum album, Seseli osseum, Festuca pallens, and Jovibarba sobolifera are indicated as diagnostic for $4-5$ vegetation units in Table 6 but for no single unit in

Table 5. These results suggest that for determination of diagnostic species generally valid over a large area, fidelity calculation in a large data set is a promising alternative to expert knowledge.

## Effects of data set structure on diagnostic species

A conspicuous discrepancy is found in vegetation units with relatively many relevés ( $\mathrm{A}, \mathrm{B}, \mathrm{E}$ and F ), which have few diagnostic species if considered in a small data set (Table 5) but several diagnostic species if considered in the large data set with additional relevés (Table 6). Numbers of diagnostic species between the two tables are directly comparable because each table contains a total of 50 diagnostic species.

Let us consider vegetation unit A (204 relevés) in a data set of 502 relevés (Table 5). In this small data set, vegetation unit A is relatively large. To be diagnostic for this unit, a species must be rather common in the data set and most of its occurrences must be concentrated in the vegetation unit. As most species are relatively rare in phytosociological data sets, very few species exceed the threshold fidelity value.

In the data set of 15989 relevés, vegetation unit A has the same absolute size but becomes small relative to the data set size. As the additional relevés belong to other vegetation types, there are several species which occur in the vegetation unit but are quite rare or absent in the additional relevés. Thus, relative concentration of these species in the vegetation unit increases and they eventually exceed the threshold fidelity value.

In conclusion, it is necessary to realize that the diagnostic capacity of species can be underestimated in vegetation units which are large relative to the data set size. However, this may be a problem only of using small data sets, because in large data sets with additional relevés all vegetation units are relatively small.

## Well defined and poorly defined vegetation units

When evaluating classifications, phytosociologists often think in terms of 'good' and 'bad' vegetation units (Pignatti et al. 1995). Vegetation units containing several faithful species are usually considered as 'good' and the others as 'poor'. 'Goodness' of a vegetation unit can be quantified by calculating the average positive fidelity

Table 7. Averages of positive values of the phi coefficient $(\times 1000)$ for all species occurring in vegetation units from Tables 5 and 6.

| Vegetation unit | A | B | C | D | E | F | G | H |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Table 5 (small) | 86 | 81 | 193 | 209 | 96 | 104 | 241 | 228 |
| Table 6 (large) | 57 | 38 | 54 | 88 | 48 | 52 | 95 | 47 |

value for all species occurring in this vegetation unit (Table 7). Vegetation units C, D, G, and H appear as the 'best' in the small data set (Table 5) but if the large data set with additional relevés (Table 6) is considered, it becomes clear that only units D and G are really 'good', while H is rather 'poor'. This indicates that many species which appeared to be faithful to communities C and H in the small data set have their main ecological optimum outside this data set, in vegetation types other than the rock-outcrop grasslands. Vegetation unit $B$ is indicated as the 'poorest' in both tables; this reflects the fact that it is mainly defined by the absence of diagnostic species of other units, combined with dominance of a single species, Allium montanum.

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