Catalogue of alien plants of the Czech Republic (2nd edition): checklist update, taxonomic diversity and invasion patterns


A complete list of all alien taxa ever recorded in the flora of the Czech Republic is presented as an update of the original checklist published in 2002. New data accumulated in the last decade are incorporated and the listing and status of some taxa are reassessed based on improved knowledge. Alien flora of the Czech Republic consists of 1454 taxa listed with information on their taxonomic position, life history, geographic origin (or mode of origin, distinguishing anecophyte and hybrid), invasive status (casual; naturalized but not invasive; invasive), residence time status (archaeophyte vs neophyte), mode of introduction into the country (accidental, deliberate), and date of the first record. Additional information on species performance that was not part of the previous catalogue, i.e. on the width of species’ habitat niches, their dominance in invaded communities, and impact, is provided. The Czech alien flora consists of 350 (24.1%) archaeophytes and 1104 (75.9%) neophytes. The increase in the total number of taxa compared to the previous catalogue (1378) is due to addition of 151 taxa and removal of 75 (39 archaeophytes and 36 neophytes), important part of the latter being the reclassification of 41 taxa as native, mostly based on archaeobotanical evidence. The additions represent taxa newly recorded since 2002 and reported in the national literature; taxa resulting from investigation of sources omitted while preparing the previous catalogue; redetermination of previously reported taxa; reassessment of some taxa traditionally considered native for which the evidence suggests the opposite; and inclusion of intraspecific taxa previously not recognized in the flora. There are 44 taxa on the list that are reported in the present study for the first time as aliens introduced to the Czech Republic or escaped from cultivation: Abies concolor, A. grandis, A. nordmanniana, Avena sterilis subsp. ludoviciana, A. xvilis, Berberis julianae, B. thunbergii, Bidens ferulifolia, Buddleja alternifolia, Buglossoides incrassatus subsp. spilitgerberi, Buxus sempervirens, Corispermum declinatum, Cotoneaster dielsianus, C. divaricatus, Euphorbia myrsinites, Gleditsia triacanthos, Helleborus orientalis, Hieracium heldreichii, Koelreuteria paniculata, Lonicera periclymenum, Lotus ornithopodioides, Malus baccata, M. pumila, Miscanthus sacchariflorus, Morus alba, Muscari armeniacum, Paeonia lactiflora, Pennisetum alopecuroides, Pinguicula crystallina subsp. hirtiflora, P. grandiflora subsp. rosea, Podophyllum hexandrum, Pyracantha coccinea, Rhodotypos scandens, Rumex patientia & R. tianschanicus ‘Uteuša’, Salix cordata, Sarracenia purpurea, Sasa palmata ‘Nebulosa’, Scolymus maculatus, Spiraea japonica,
Tagetes tenuifolia, Thuja occidentalis, Trifolium badium, Vaccinium corymbosum and Viburnum rhytidophyllum. All added and deleted taxa are commented on. Of the total number of taxa, 985 are classified as casuals, 408 as naturalized but not invasive, and 61 as invasive. The reduction in the number of invasive taxa compared to the previous catalogue is due to a more conservative approach adopted here; only taxa that currently spread are considered invasive. Casual taxa are strongly over-represented among neophytes compared to archaeophytes (76.7% vs 39.4%), while naturalized but non-invasive taxa follow the reversed pattern (18.8% vs 57.4%). However, these two groups do not significantly differ in the proportion of invasive taxa. Of introduced neophytes, 250 taxa (22.6%) are considered vanished, i.e. no longer present in the flora, while 23.3% became naturalized, and 4.5% invasive. In addition to the traditional classification based on introduction–naturalization–invasion continuum, taxa were classified into 18 population groups based on their long-term trends in metapopulation dynamics in the country, current state of their populations, and link to the propagule pressure from cultivation. Mapping these population groups onto the unified framework for biological invasions introduced by Blackburn et al. in 2011 made it possible to quantify invasion failures, and boom-and-busts, in the Czech alien flora. Depending on inclusion criteria (whether or not extinct/vanished taxa and hybrids are considered), alien taxa ever recorded in the Czech Republic contribute 29.7–33.1% to the total country’s plant diversity; taking into account only naturalized taxa, a permanent element of the country’s flora, the figure is 14.4–17.5%. Analysis of the dates of the first record, known for 771 neophytes, indicates that alien taxa in the flora have been increasing at a steady pace without any distinct deceleration trend; by extrapolating this data to all 1104 neophytes recorded it is predicted that the projected number would reach 1264 in 2050. Deliberate introduction was involved in 747 cases (51.4%), the remaining 48.6% of taxa are assumed to have arrived by unintentional pathways. Archaeophytes are more abundant in landscapes, occupy on average a wider range of habitat types than neophytes, but reach a lower cover in plant communities. The alien flora is further analysed with respect to representation of genera and families, origin and life history.

Keywords: abundance, alien flora, checklist, casual, cover in plant communities, Czech Republic, exotic species, geographic origin, habitat niche, hybridization, impact, introduction–naturalization–invasion continuum, invasive plants, life history, naturalized, non-native species, residence time, taxonomy

Introduction

The last decade was a period of intensive research on biological invasions in Europe (see Pyšek & Hulme 2011 for review), an important part of which represented the collation of regional data on alien plant species. With the exception of the UK (Clement & Foster 1994, Ryves et al. 1996, Preston et al. 2002), complete checklists of alien floras for European countries only started to appear at the beginning of the 2000s (Essl & Rabitsch 2002, Klotz et al. 2002, Reynolds 2002). The first comprehensive checklist of alien plants in the Czech Republic was published 10 years ago as a part of the Catalogue of alien plants of the Czech Republic (Pyšek et al. 2002). It provided information on 1378 alien taxa and stimulated development of the associated database CzechFlor, held at the Institute of Botany AS CR in Průhonice. These data, together with other datasets resulting from recent research, have been used for a number of analyses of plant invasions in the country that addressed issues such as species invasiveness (Kubešová et al. 2010, Moravcová et al. 2010), associations with pollinators (Pyšek et al. 2011a), habitat invasibility (Chytrý et al. 2005, 2008a, 2009b, Sádlo et al. 2007), rates of spread and range filling (Williamson et al. 2005, 2009, Pyšek et al. 2011c), interaction of traits, propagule pressure and residence time in affecting invasion success (Pyšek et al. 2009b), pathway efficiency (Pyšek et al. 2011b), and risk assessment (Křivánek & Pyšek 2006, Chytrý et al. 2009b). In addition, data on native
species that are also part of the CzechFlor database provided basis for analyses of the performance of central-European species as aliens in other parts of the world (Pyšek et al. 2009a, Phillips et al. 2010, Stohlgren et al. 2011). Within the DAISIE and ALARM (Settele et al. 2005) projects, the data from the 2002 catalogue were part of the pan-European dataset that was used to analyse invasion patterns at the continental level, including cross-taxonomic evaluation of the role of macroeconomic and demographic factors in determining regional levels of invasion (Pyšek et al. 2010b, Essl et al. 2011), distribution of alien species in habitats (Pyšek et al. 2010a), assessment of ecological and economic impacts of alien species in Europe (Winter et al. 2009, Vilà et al. 2010) and risk-assessment for plants based on habitat mapping (Chytrý et al. 2008b, 2009a, 2012).

These studies clearly indicate the value of complete national or regional checklists for understanding invasions. This started to be fully recognized in the 2000s and resulted in a call for pan-European inventory of invasive species within the European framework programmes; until then there was some information on alien floras available for European countries (Weber 1997), but the quality of data was highly variable (Pyšek 2003). The DAISIE project (2004–2008) made it possible to organize and develop this line of research based on extensive international cooperation in Europe (DAISIE 2009). The project assembled available data on alien plants for 48 European countries and regions, which until then were scattered in a variety of published and unpublished accounts and databases. For some countries DAISIE collected the first comprehensive checklists of alien species based on primary data (Lambdon et al. 2008), and established an online database, the European Invasive Alien Species Gateway (DAISIE 2008). At the same time it stimulated elaboration of comprehensive alien species checklists in individual countries, a process that still continues, and yielded new plant data for e.g. Belgium (Verloove 2006), Estonia (Ööpik et al. 2008), Italy (Celesti-Grapow et al. 2009), Greece (Arianoutsou et al. 2010), and most recently Slovakia (Medvecká et al. 2012).

The Czech Republic, a central-European country with an area 78,864 km², 10.3 million inhabitants, and a human population density of 131 inhabitants per km², is prone to plant invasions due to historical and geographical factors: location on the crossroads of the continent, many natural or human-created migration routes opening possibilities for colonization, and long-lasting human influence that further diversified the naturally diverse and heterogeneous landscape mosaic (see Pyšek et al. 2002 for details). These features, together with a strong botanical tradition and in-depth knowledge of plant communities (Chytrý 2007, 2009, 2011) make the country a suitable model for studying regional patterns of plant invasions. In the last decade since the publication of the previous catalogue a wealth of information on alien species has been accumulated, which created a need for a revision of the original checklist.

The aim of the present paper is to update and improve the original checklist of alien plant taxa in the Czech Republic (Pyšek et al. 2002) by incorporating new data accumulated in the last decade, reassessing the status of taxa resulting from improved taxonomic knowledge, and wherever needed, correcting errors which can hardly be avoided in such a comprehensive work. We also provide additional information that was not part of the previous catalogue, including the width of species’ habitat niches, their dominance in invaded communities and their impacts. Changes from the 2002 version are documented so that the reasoning behind them can be followed.
Methods

Data sources

The basis for the present checklist was the Catalogue of alien plants of the Czech Republic published a decade ago (Pyšek et al. 2002). For historical data, the compilation of both the previous and current checklist relied on an outstanding tradition of the floristic research in the Czech Republic dating back to the second half of the 18th century (reviewed in detail in Pyšek et al. 2002). Already in the 19th century, a series of floras and species lists were published, covering the present territory of the Czech Republic (see Krahulec 2012 for a review of the history of botanical research), and recognizing plants by geographic origin; these provide valuable information about the occurrence of plants at those times and residence times of neophytes (Pohl 1809–1814, Presl & Presl 1819, Opiz 1823, 1852, Rohrer & Mayer 1835, Makowsky 1863, Oborny 1886, Formánek 1887–1897). The wealth of information on alien plants can be found especially in the remarkable works by Čelakovský (1868–1883, 1882–1894), who recognized the alien status and origin of some plants present in the Czech flora and commented in considerable detail on their distribution. The recognition of alien plants continued in floras and specialized studies in the 20th century (e.g. Polívka 1900–1904, Laus 1908, Domin 1917, 1918, 1919, Dostál et al. 1948–1950, 1954, 1958, 1989). Since the 1960s, systematic attention started to be paid to plants, including aliens, in specific human-made habitats (ports, railways, oilseed or wool processing factories, grain silos, mills, rubbish tips, arable land, etc.) thanks to a specialized research section established at the Institute of Botany, Průhonice, in the 1960s. This work yielded several focused compendia (e.g. Hejný et al. 1973) and provided a basis for systematic recording of alien plants (e.g. Jehlík 1986, 1998a).

The Flora of the Czech Republic, with eight of nine planned volumes published up to now (Hejný & Slavík 1988–1992, Slavík 1995, 1997a, 2000, Slavík & Štěpánková 2004, Štěpánková 2010) and the Key to the flora of the Czech Republic (Kubát et al. 2002), served as a fundamental information source for this checklist. Other recent sources included national floristic literature, namely that published in the journals of the Czech Botanical Society (see References). During the last decade, new records for the flora of the Czech Republic have been systematically reported in an annually published series, Additamenta ad floram Reipublicae Bohemicae, which has thus far yielded 10 summarizing accounts (Hadinec et al. 2002, 2003, 2004, 2005, Hadinec & Lustyk 2006, 2007, 2008, 2009, 2011, 2012). The series, initiated and edited by J. Hadinec, in cooperation with František Procházka and Pavel Lustyk, proved a valuable source because it not only reports new finds but also critically re-evaluates status of particular species and provides additional data on their distribution.

For archaeophytes, a strong tradition of Czech archaeobotanical research provided a solid basis for evaluation of species origin and immigration status. Main sources include the works of E. Opravil and V. Čulíková (see References), the results of which are now available in the Archaeobotanical database of the Czech Republic (CZAD; Archaeological Institute AS CR 2011).

Other data sources included unpublished information provided by many colleagues (see Acknowledgments), herbarium collections to verify some literature reports (namely PR, PRC, BRNU and PRA; codes follow Thiers 2012) and our own floristic field records from 2002–2012.
The data presented here and in the previous catalogue (Pyšek et al. 2002) are organized in the working database CzechFlor held at the Institute of Botany AS CR, Průhonice.

Classification of taxa: invasion status

This work focuses on alien species (synonyms: adventive, exotic, introduced, non-indigenous, non-native) in the Czech Republic which we define as species present in the region because human actions enabled them to overcome fundamental biogeographical barriers (i.e. human-mediated extra-range dispersal); they occur in the area as a result of intentional or accidental introduction by humans, or of a spontaneous spread from other regions where they were introduced by humans. Crosses resulting from hybridization with one or both alien species involved are considered alien (Pyšek et al. 2004a). We define native species (synonym: indigenous species) as those that have evolved in a given area or that arrived there by natural means (through range expansion) without any intentional or accidental intervention of humans from an area where they are native (Pyšek et al. 2004a).

We classified species according to the stage they reached along the introduction–naturalization–invasion continuum (INIC) that describes how species proceed in the invasion process by overcoming geographical, environmental and biotic barriers (Richardson et al. 2000, 2011, Richardson & Pyšek 2006, Blackburn et al. 2011). Based on this concept we use the following terms to describe the invasion status:

(i) Casual species are those alien species that do not form self-sustaining populations in the invaded region; they may flourish and reproduce occasionally in an area but their persistence depends on repeated introductions of propagules.

(ii) Naturalized species (synonym: established species) form self-sustaining populations for several life cycles without direct intervention by people, or despite human intervention; they often recruit offspring freely, usually close to adult plants and their persistence does not depend on ongoing input of propagules.

(iii) Invasive species are a subset of naturalized species; they form self-replacing populations over many life cycles, produce reproductive offspring, often in very large numbers at considerable distances from the parent and/or site of introduction, and have the potential to spread over long distances. In addition to this definition, we introduce the metapopulation criterion to separate invasive species from naturalized, to account for the historical population dynamics of the treated taxa (see the next section).

We included in the list all taxa that were reported to occur at least once in the wild, while those kept exclusively in cultivation are not considered. For escapees from cultivation, a plant was included in the list if it reproduced on its own outside the space where it was sown or planted (Pyšek et al. 2002). In plants reproducing by seed, germination outside such space was considered as an escape from cultivation. A plant reproducing clonally was considered as an escape from cultivation only if it survived winter and persisted in a given site until the following growing season.

Compared to the previous catalogue (Pyšek et al. 2002), we adopted a more conservative approach; if there were doubts about a species’ origin status and no strong evidence to consider it alien, it was not included in the list; this conservative approach resulted in removing some species that were listed in the previous catalogue (see Appendix 1).

The classification of casual vs naturalized status is especially difficult for woody plants reproducing in the parks or gardens where they are planted; in some cases this happens...
over a large area and for decades (e.g. many trees and shrubs in the Průhonice Park near Prague where there is a long-term systematic recording of regeneration). Here we aimed at adopting the criterion of reproduction over several generations (Richardson et al. 2000) which puts the time criterion in a different perspective than that applied for non-woody taxa. Such taxa are therefore mostly classified as casual. Also, the majority of hybrids are considered casual, with the exception of stabilized hybrids that include some naturalized (e.g. *Medicago × varia*, *Helianthus × laetiflorus*, *Mentha × rotundifolia* and *Oenothera spp.*) or invasive taxa (e.g. *Reynoutria × bohemica*, *Populus × canadensis* and *Symphyotrichum × versicolor*).

Unlike the previous catalogue (see Pyšek et al. 2002 and their Appendix 1), we do not explicitly label taxa as locally naturalized. In the present paper this can be inferred from the combination of invasion status and regional abundance category in Appendix 2. In the same vein, taxa are not labelled as post-invasive since this status is included in the classification using the population groups (see below).

**Classification of taxa into categories based on long-term population dynamics and historical link with cultivation: incorporating the unified framework for biological invasions**

In addition to traditional classification scheme dividing species into three basic categories along the INIC (Richardson et al. 2000, Richardson & Pyšek 2006, Pyšek & Richardson 2010) here we attempt for an even finer classification based on the population approach emphasized by Blackburn et al. (2011). The basis for this classification are the criteria of reproduction and survival applied against the background of the metapopulation approach. This makes it possible to separate species that survive in a single or few populations in a spatially restricted area from those that spread and form metapopulations over large areas.

Another important point to emphasize is that we refer to the population history viewed from the **current perspective**, i.e. the state in which the populations of a given species exist at present. Therefore, invasions that proved unsuccessful in proceeding along the various stages of the INIC (see Blackburn et al. 2011 and their Fig. 1) are reflected in the current classification, and in changes of invasion status compared to the previous treatment (Pyšek et al. 2002). From this it follows that some taxa that were previously classified as naturalized are moved to the casual category (reflecting ‘invasion failure’), and some taxa previously considered invasive are now classified as naturalized (reflecting ‘boom and bust phenomenon’; sensu Blackburn et al. 2011). These shifts among the INIC categories reflect not only changes in species’ behaviour in the past decade but also the more conservative approach adopted for the current classification. Another principle we follow is that of the **highest stage achieved** at the population level; individual populations of an alien species may occur in a region in different stages of the INIC; early in the process, some can be naturalized while others are still casual (e.g. Essl et al. 2009), whereas later on, some can be invasive while others not (e.g. Meyerson et al. 2010a, b, Saltonstall et al. 2010). Therefore, if some of the populations of a species reached the naturalized or invasion stage, the species is classified as such in Appendix 2.

Therefore, the rationale of classification of alien species into finer groups (termed ‘population groups’) is based on the following criteria (Table 1):
(A) **Sustainability of populations** of the species in the target region of the Czech Republic; here we distinguish between (i) species existing as non-self-sustaining populations or occasionally recorded individuals, corresponding to Blackburn et al.’s (2011) categories B3+C2, and the casual stage of Richardson et al.’s (2000) framework; the reason for lumping the categories B3 (defined as individuals transported beyond limits of native range, and directly released into novel environment) and C2 (individuals surviving in the wild in location where introduced, reproduction occurring, but population not self-sustaining) is that from records in floristic literature it is impossible to infer whether the presence of the plant is due to a direct introduction of a propagule into the region or a result of a temporary reproductive event within the region; (ii) species occurring in self-sustaining populations; these populations can be numerous and widespread but remain isolated (C3+D1+D2, naturalized species – lumping due to insufficient knowledge about whether the populations recruit from the original point of introduction and whether those spread far from it reproduce in new locations); and (iii) species that currently form numerous and persistent metapopulations widespread over large areas (Blackburn et al.’s 2011 category E).

(B) **Historical population dynamics** is used to classify species according to the highest stage they reached in the invasion process combined with the current state. We distinguish whether or not the most successful populations of unsuccessful species have established and were surviving in the region before decline to the current levels of occurrence; successful species are classified based on the tendency for spread, with respect to whether this trend occurred in the past or is still valid (Table 1). Employing this criterion, i.e. focus on the current status of species’ populations and processes that resulted in the present state, is the reason why the correspondence with the categories of Blackburn et al. (2011) is, however, not automatically translated into those of the introduction–naturalization–invasion continuum. This concerns those species classified as D1, D2 and considered invasive in Blackburn et al.’s (2011) scheme (self-sustaining population in the wild, with individuals surviving, or also reproducing, a significant distance from the original point of introduction), populations of which no longer exhibit dynamic spread and are currently stabilized (Groups 7, 9, 11 in Table 1), or even decline in the Czech Republic (Group 6). We also do not consider as invasive those species that only start to exhibit symptoms of the beginning spread (Groups 8, 10, 12). Adhering to a conservative approach, these species are still considered as naturalized. Nevertheless, they merit particular attention in terms of monitoring as they are likely to become invasive in the near future. Only those species that are currently spreading are classified as invasive (Groups 14, 16, 18; Table 1).

(C) **Link to populations in cultivation**. The above criteria are employed against the background of species’ planting histories in the region. Here we separate species into (i) those that have never been cultivated (corresponding to contaminant and stowaway pathways of introduction according to Hulme et al. 2008; Appendix 2), hence unsupported by the propagule pressure from planted populations; (ii) those in which the peak of planting intensity was in the past and at present the planting ceased or is only of marginal importance; and (iii) those that are still commonly kept in cultivation, be it for horticultural or agricultural purposes. For the cultivated species this criterion refers to the degree of continuity of propagule pressure. The time frame over which this criterion applies is the last ca 200 years for which period the information on the frequency of planting can be inferred.
Table 1. – Classification of the alien flora of the Czech Republic into population groups (PG) based on the current population state and their connectivity, trends in their long-term dynamics, and link to cultivated populations as a source of propagule pressure in the past and present. See text for details. The population groups are referred by numbers presented in Appendix 2, with the INIC (introduction–naturalization–invasion continuum) status indicated and number of species shown in parentheses. The link to the unified invasion framework (Blackburn et al. 2011) is indicated by their categories that are relevant to the given population state shown in parentheses; note that some of their categories referring to the invasion stage such as D1, D2, E (Blackburn et al. 2011; their Fig. 1) are classified as naturalized because the focus here is on the present state and approach adopted is conservative. Taxa in these categories may have reached the invasion stage in the past but their populations are stabilized and no longer spread. Link to standard classification of the INIC categories (Richardson et al. 2000) is indicated by coloured shading. The scheme also separates groups of taxa introduced by unintentional pathways (contaminant, stowaway), marked “none” in the Cultivation column, from those introduced deliberately (release, escape; Hulme et al. 2008, Pyšek et al. 2011b).

<table>
<thead>
<tr>
<th>Populations</th>
<th>Cultivation</th>
<th>Introduction &amp; Failure</th>
<th>Establishment &amp; Failure</th>
<th>Establishment &amp; No trend</th>
<th>Starting spread</th>
<th>Ongoing spread</th>
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<tbody>
<tr>
<td>(a) Not self-sustaining</td>
<td>(a1) None</td>
<td>PG1: casual (395)</td>
<td>PG2: casual (45)</td>
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<tr>
<td>(B3, C2)</td>
<td>(a2) Past</td>
<td></td>
<td>PG3: casual (17)</td>
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<td></td>
<td>(a3) Ongoing</td>
<td>PG4 &amp; 5: casual (501 &amp; 28)</td>
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<tr>
<td>(b) Self-sustaining</td>
<td>(b1) None</td>
<td>PG6: naturalized (54)</td>
<td>PG7: naturalized (40)</td>
<td>PG8: naturalized (43)</td>
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<tr>
<td>(C3, D1, D2)</td>
<td>(b2) Past</td>
<td></td>
<td>PG9: naturalized (36)</td>
<td>PG10: naturalized (11)</td>
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<td></td>
<td>(b3) Ongoing</td>
<td></td>
<td>PG11: naturalized (65)</td>
<td>PG12: naturalized (31)</td>
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<tr>
<td>(c) Metapopulations</td>
<td>(c1) None</td>
<td></td>
<td>PG13: naturalized (100)</td>
<td></td>
<td>PG14: invasive (28)</td>
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<tr>
<td>(E)</td>
<td>(c2) Past</td>
<td>PG15: naturalized (8)</td>
<td></td>
<td>PG16: Invasive (9)</td>
<td></td>
<td></td>
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<tr>
<td></td>
<td>(c3) Ongoing</td>
<td>Group 17: naturalized (19)</td>
<td></td>
<td>PG18: invasive (24)</td>
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<td>Total taxa</td>
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<td></td>
<td></td>
<td>924</td>
<td>116</td>
<td>268</td>
<td>85</td>
<td>61</td>
</tr>
</tbody>
</table>
Residence time status

Based on the residence time, i.e. the time since the arrival of a species to the territory of the present Czech Republic, we distinguish archaeophytes (taxa introduced before the discovery of America, approx. 1500 A. D.) and neophytes (taxa introduced after that date), following the concept traditionally used in European studies on plant invasions (e.g. Holub & Jirásek 1967, Pyšek et al. 2002, 2004a). When evaluating residence time status of hybrids, we followed that of the alien parent; therefore, crosses of archaeophytes with native are considered archaeophytes, and hybridization with neophytes involved are classified as neophytes regardless of the status of the second parent.

For neophytes, we determined the year of the first record in the Czech Republic that is used to infer the minimum residence time, i.e. the time for which the species is known to be present (Rejmánek 2000, Pyšek & Jarošík 2005, Richardson & Pyšek 2006); this characteristic is important in evaluation of invasion status since it indicates how much time the species had to colonize suitable habitats (Williamson et al. 2009, Gassó et al. 2010), go through a lag phase (Kowarik 1995, Crooks 2005) or build relationship with native biota (Pyšek et al. 2011a). As pointed out above, the reliability of the years of first records crucially depends on the intensity of floristic research in the past (see Pyšek et al. 2002 for discussion).

Species traits: taxonomic affiliation and life history

Taxonomic affiliation of taxa to families follows the approach of the Angiosperm Phylogeny Group Classification: APG III (Stevens 2001 onwards, Angiosperm Phylogeny Group 2009), and Smith et al. (2006) for ferns. This classification system incorporates data from molecular, chemical and morphological phylogenies in an attempt to represent the latest thinking on angiosperm evolution, and in a few lineages (e.g. Scrophulariales) it differs markedly from the traditional system.

The following life histories were assigned to the species: annual, biennial, perennial, semishrub, shrub, tree, fern, aquatic and parasitic (see Appendix 2).

Geographic origin

Taxa were classified according to their geographic origin (native range) at the level of continents (parts of Europe other than the Czech Republic, Africa, Asia, North America including Mexico, Central America, South America, and Australia). Unlike the previous catalogue (Pyšek et al. 2002), we distinguished the Mediterranean region as a separate region of origin, covering respective parts of southern Europe, northern Africa and western Asia from Turkey and Israel to Afghanistan. This broad definition of the Mediterranean region corresponds to the Mediterranean, Submediterranean and Oriental Floristic Subregions according to Meusel et al. (1965). The region delimited in this way is very convenient for plant invasion studies as it includes the areas of origin of Neolithic agriculture. Indications of Europe, Asia and Africa in Appendix 2 refer to their parts other than the Mediterranean region in this delimitation.

Hybrids and species that originated through recent hybridization are listed as a special origin category and we employed classification based on how species originated in terms of their evolutionary history. This approach acknowledges that some did not evolve naturally, but under human influence, do not have a natural home range, and their original hab-
itat is unknown (Kühn & Klotz 2003). Especially for many archaeophytes, native ranges are not known or are highly uncertain, and some archaeophytes are regarded as alien throughout their known global range. These taxa, termed anecophytes (homeless plants; Zohary 1962) could be cultivated plants that escaped to the wild or plants that co-evolved with human land uses such as agriculture (Kühn & Klotz 2002, 2003, Kühn et al. 2004). In our treatment, we follow the more conservative approach and label as anecophytes mostly those species that evolved in cultivation, or species occurring in the wild but with their region of origin being unknown.

**Regional abundance**

Type of regional abundance in the landscape was estimated for each taxon using the following scale: single locality, rare, scattered, locally abundant, and common across the whole Czech Republic. A special category termed ‘vanished’ relates to the taxa for which no records have been known for a long period, and where it is highly improbable that they would appear again (Pyšek et al. 2002).

**Occurrence in habitats**

The previous catalogue provided information on the occurrence of alien species in phytosociological alliances, different types of landscapes and with respect to landuse (Pyšek et al. 2002). Here we use extensively revised data from the database of species occurrences in 88 major habitat types of the Czech Republic as defined by Sádlo et al. (2007), which correspond to phytosociological alliances or groups of alliances. All four levels of species affinity to the habitats as defined by Sádlo et al. (2007: 305) are taken into account, i.e. a species is considered as occurring in a habitat even if the habitat is outside its ecological optimum, but the species is occasionally found there.

**Cover in plant communities**

To obtain the data on the cover of alien species in plant communities, we used vegetation plot observations (phytosociological relevés) stored in the Czech National Phytosociological Database held at the Department of Botany and Zoology, Masaryk University, Brno (Chytrý & Rafajová 2003, EU-CZ-001 according to Dengler et al. 2011). At the time of data extraction (April 2012) the database contained 88,215 relevés from plots smaller than 1000 m² with an indication of plot size and geographical coordinates. Of these, 41,582 relevés contained at least one alien species. To reduce oversampling of some areas or some vegetation types, we selected only one relevé from a group or relevés assigned to the same phytosociological alliance within the same grid cell of 1.25 longitudinal × 0.75 latitudinal minutes, i.e. approximately 1.5 × 1.4 km. This stratified resampling yielded 16,033 relevés containing 437 alien species, which were used to quantify species cover. Only species occurring in at least 25 relevés were evaluated to avoid inaccuracies resulting from small sample size. For these species, mean percentage cover across all relevés in which the species was present was calculated.

**Impact**

To provide the first insights into the impacts of alien plant species in the Czech Republic, we used the data gathered by the DAISIE project (DAISIE 2008, 2009) and indicated
those species on our list for which an ecological and/or economic impact is reported in the literature (Vilà et al. 2010). With a few exceptions indicated in Appendix 2, this classification has not been done specifically for the Czech Republic but refers to any region in Europe, meaning that species labelled as exerting impact may not do so in this country.

**Statistical analysis**

To test whether there are differences between species numbers according to their invasion status, life histories, abundances and origins, their counts were analysed by row × column contingency tables, using generalized linear models with log-link function and Poisson distribution of errors (e.g. Crawley 1993: 231–237). To ascertain for which species the counts are lower or higher than would be expected by chance, adjusted standardized residuals of G-tests were compared with critical values of normal distribution (Řehák & Řeháková 1986). The estimates of yearly accumulations of neophytes, including projected total numbers in 2050, were assessed from linear regressions of cumulative numbers that started in the year 1800.

**Results and discussion**

*Diversity of alien flora*

The alien flora of the Czech Republic consists of 1454 taxa, made up by 350 archaeophytes (24.1%) and 1104 neophytes (75.9%; Table 2, Appendix 2), which represent addition to ca 2945 native taxa known from the country (using a preliminary estimate from Danihelka et al. 2012) and form 33.1% of the total plant diversity ever recorded there. Although similar figures for individual countries are subject to variation resulting not only from composition of floras but also from the variable depth of their knowledge, intensity of research into alien species, or whether apomictic species are included in comparisons (see Williamson 2002, Pyšek et al. 2002 and discussion therein), the proportion given here seems to reasonably reflect situation in countries with detailed knowledge of their floras. Subtracting species that are assumed to be vanished among alien (277 taxa, Appendix 2) and extinct from native flora (153 taxa in the Red List categories A1 and A2; Danihelka et al. 2012) yields a figure of 29.7% of aliens contributing to the plant diversity currently occurring in the Czech Republic.

Table 2. – Numbers of all alien taxa in the Czech Republic, including hybrids, cross-tabulated across invasion status and immigration time. Note that invasive taxa are subgroup of naturalized. Overall, the observed counts of alien taxa (in bold) highly significantly ($\chi^2 = 193.56; \text{df} = 2; P < 0.0001$) differ from counts expected by chance (values in parentheses). Statistically highly significant deviations of individual counts from counts that can be expected by chance are expressed by asterisks (*** $P < 0.001$); numbers in parentheses not followed by any symbol do not differ from randomly expected values.

<table>
<thead>
<tr>
<th></th>
<th>Naturalized non-invasive</th>
<th>Invasive</th>
<th>Total</th>
</tr>
</thead>
<tbody>
<tr>
<td>Archaeophytes</td>
<td>201 (97.6)***</td>
<td>11 (14.8)</td>
<td>350</td>
</tr>
<tr>
<td>Neophytes</td>
<td>207 (310.3)***</td>
<td>50 (47.2)</td>
<td>1104</td>
</tr>
<tr>
<td>All aliens</td>
<td>408</td>
<td>61</td>
<td>1454</td>
</tr>
</tbody>
</table>
If we further exclude 94 hybrids recorded from the total number of alien taxa, and compare this figure with the current native species diversity without 575 hybrids (Danihelka et al. 2002), the proportion of alien taxa 32.8%. The hybrids between neophytes and native taxa, and between two neophytes, are more frequent than hybrids involving archaeophytes. Overall, neophytes are involved in 58 hybrid combinations, archaeophytes in 42 and native species in 56 (Table 3).

Finally, considering only permanently present taxa, i.e. 469 naturalized aliens (including both non-invasive and invasive) and the native flora without extinct representatives, yields 14.4% contribution of alien flora to the current plant diversity, or 17.5% if hybrids are excluded from native flora. This proportion is probably a more realistic measure of the level of invasion of the country’s species pool than is usually given in overall figures based on all species ever recorded, including casuals, because it better reflects the threat from alien species’ impacts and potential for invasion debt to operate (Essl et al. 2011).

Table 3. – Numbers of hybrids in the alien flora classified according to the origin and residence time status of their parental species. Note that the total number of hybrids across the three groups (n = 94) does not correspond to the sum of numbers within the groups involved because all combinations are displayed row-wise. Anecophytes are listed as species of unknown origin, the majority of which originated by hybridization in cultivation. Hybrids of native species are not relevant (n.r.) for this comparison.

<table>
<thead>
<tr>
<th>Total within group</th>
<th>× Archaeophyte</th>
<th>× Neophyte</th>
<th>× Native</th>
</tr>
</thead>
<tbody>
<tr>
<td>Archaeophyte</td>
<td>13</td>
<td>6</td>
<td>23</td>
</tr>
<tr>
<td>Neophyte</td>
<td>6</td>
<td>19</td>
<td>33</td>
</tr>
<tr>
<td>Native</td>
<td>23</td>
<td>33</td>
<td>n.r.</td>
</tr>
<tr>
<td>Hybrids total</td>
<td>94</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Anecophytes</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Hybrids and anecophytes total</td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

Changes to the 2002 checklist

Compared to the first checklist (Pyšek et al. 2002), 75 taxa were removed (39 archaeophytes and 36 neophytes). The majority of these changes resulted from reclassifying some taxa as native (41 taxa) where evidence for their alien origin was not convincing enough under the conservative approach adopted in the present paper; they were mostly archaeophytes but there are also six neophytes with alien status which appeared doubtful based on recently published evidence: Agropyron pectinatum, Crocus heuffelianus, Epilobium dodonaei, Senecio rupestris, Teucrium scorodonia and Viola tricolor subsp. curtisii. For nine taxa previously classified as deliberately introduced casuals, the evidence for escaping from cultivation was ambiguous. Other deletions relate to 10 taxonomically unjustified taxa now omitted from the Czech flora, and 16 cases are doubtful records previously only reported in the literature that cannot be considered as proven without herbarium evidence, or taxa that were erroneously determined by the collector. All deleted species are dealt with in detail in Appendix 1.

In total, 151 taxa not listed in Pyšek et al. (2002) are included, representing additions to the alien flora of the Czech Republic. This includes taxa newly recorded since 2002 and (i) reported in the literature (e.g. Convallaria majalis var. transcaucasica, Darmera peltata,
Dittrichia graveolens, Euphorbia agraria, Galium murale, Geranium purpureum, Gratiola neglecta, Hypericum annulatum, Legousia pentagonia, Pimpinella peregrina and Stachys setifera), including two volumes of the Flora of the Czech Republic published in this period (Slavík & Štěpánková 2004, Štěpánková 2010) that report taxa missing from previous catalogue (e.g. Cichorium endivia, Egeria densa and Filago pyramidata); (ii) additions resulting from investigation of sources omitted from the previous catalogue (e.g. Euphrasia salisburgensis, Herniaria incana, Rumex longifolius subsp. sourekii, Trifolium badium and Xerocrysum bracteatum), including some herbarium materials (e.g. Centaurea carniolica, C. transalpina and Corispermum declinatum); (iii) redetermination of previously reported taxa (e.g. Eriochloa punctata, Gilia achilleifolia, Hieracium sp. ex H. heldreichii agg., Rodgersia pinnata and Spiraea hypericifolia subsp. obovata); (iv) reassessment of some taxa traditionally considered native for which the evidence suggests the opposite (e.g. Eragrostis pilosa, Lathyrus hirsutus, Lilium bulbiferum, Matricaria chamomilla and Sorbus austriaca); (v) intraspecific taxa previously not recognized in the flora (e.g. Avena sterilis subsp. ludoviciana). Accounts on the newly added alien species in the Czech flora are given in Appendix 1, with respective references.

In total, 44 taxa are reported in the present study for the first time as aliens introduced to the Czech Republic or escaping from cultivation (Appendix 1): Abies concolor, A. grandis, A. nordmanniana, Avena sterilis subsp. ludoviciana, A. ×villis, Berberis julianae, B. thunbergii, Bidents ferulifolius, Buddleja alternifolia, Buglossoides incrassata subsp. splittereri, Buxus sempervirens, Corispermum declinatum, Cotoneaster dielsianus, C. divaricatus, Euphorbia myrsinites, Gleditsia triacanthos, Helleborus orientalis, Hieracium heldreichii agg., Koelreuteria paniculata, Lonicera periclymenum, Lotus ornithopodioides, Malus baccata, M. pumila, Miscanthus sacchariflorus, Morus alba, Muscari armeniacum, Paeonia lactiflora, Pennisetum alopecuroides, Pinguicula crystallina subsp. hirtiflora, P. grandiflora subsp. rosea, Podophyllum hexandrum, Pyrananthoa coccinea, Rhodotypos scandens, Rumex patientia × R. tianschanicus ‘Uteuša’, Salix cordata, Sarracenia purpurea, Sasa palmata ‘Nebulosa’, Scolymus maculatus, Spiraea japonica, Tagetes tenuifolia, Thuja occidentalis, Trifolium badium, Vaccinium corymbosum and Viburnum rhytidophyllum.

Finally, compared to the previous version of the catalogue (Pyšek et al. 2002), 134 names were changed due to nomenclatural reasons or development in taxonomic opinion; these changes are summarized in Electronic Appendix 1.

Transitions along the introduction–naturalization–invasion continuum

Among the 1454 taxa, 985 (67.7%) are classified as casual, 408 (28.1%) as naturalized but non-invasive, and 61 (4.2%) as invasive (Fig. 1, Table 2). Among casual taxa, 86.0% are neophytes and 14.0% archaeophytes, the corresponding figures being 50.7 and 49.3%, respectively, for naturalized, and 82.0 and 18.0% for invasive taxa. From this it follows that casual taxa are strongly over-represented among neophytes, and naturalized among archaeophytes (Table 2, Fig. 1), a pattern previously illustrated for the Czech flora by Pyšek et al. (2002) and also valid for neighbouring Slovakia (Medvecká et al. 2012). Interestingly, the observed numbers of neither archaeophytes nor neophytes differ from those expected by chance, indicating that there is no difference between the two groups in the proportion of species that reach the invasion stage (Table 2, Fig. 1).
Data on neophytes provide insights into the transition rates along INIC, i.e. how large a proportion of species reach the subsequent stages of the invasion process (Fig. 2); this proportion cannot be calculated for archaeophytes because information on casual species from the initial periods of introduction is missing (Pyšek et al. 2002). Of the total number of 847 recorded casual neophytes, 250 (29.5%) have not been recorded for a long period of time and are therefore considered vanished (96 of them were only known from a single locality), and 597 (70.5%) are currently present as casuals. Of the 1104 neophytes, 257 (23.3%) became naturalized, and 50 (19.5%) of the naturalized are considered invasive (Fig. 2).

The approach we adopt takes into account invasion failures, represented by dotted arrows in Fig. 2 that indicate reversed directions in the invasion process. This makes it possible, by using finer classification based on the assessment of long-term population dynamics and its comparison with the current stage (Table 1), to map the number of taxa onto the unified framework of biological invasions (Blackburn et al. 2011). Four types of unsuccessful invasions can be recognized, depicted in Fig. 3 and based on population groups described below: (i) casual taxa that failed to establish, never forming self-sustaining populations (PG 1+4+5); (ii) taxa that formed self-sustaining populations in the past but declined so that this is no longer the case (PG 2+3); (iii) taxa present for a long time with populations surviving in the landscape; although they are still considered naturalized, their invasion obviously failed because they are rare and their decline is likely to continue (PG6); (iv) naturalized species that form stabilized metapopulations in the wild, some of them reached the invasion stage in the past but their current occurrence indicates that they declined; therefore they are considered as representatives of the boom and bust phenomenon (PG 13+15+17; Fig. 3).

Fig. 1. – Representation of taxa according to invasion status (casual, including vanished taxa; naturalized but non-invasive; invasive) among archaeophytes, neophytes and all aliens in the flora of the Czech Republic. See Table 2 for the numbers of taxa and statistics.
Overview of population groups

(a) Not self-sustaining populations or individuals
(a1) No link to cultivation

Group 1. Introduction and failure. Unintentionally introduced taxa that were only recorded as individuals or in small populations, mostly occasionally, and are reported from a single or few locations; they are classified as casuals and a significant proportion (186 of 395 in total) are considered vanished, i.e. recorded in the past and not observed for a long time since the last record. The vast majority of taxa in this group (364) are neophytes, and many occasionally recorded hybrids (75) also fall here. Typical examples include Alhagi maurorum, Chloris virgata, Cakile maritima, Conyza triloba and Scleroblitum atriplicinum.
**Group 2. Establishment and failure.** This group includes almost exclusively archaeophytes (37 of 45 in total) that were surviving in the landscape for centuries or millennia, formed self-sustained populations in the past, some of them might have been even invasive at some stage, but now they have declined or are even considered vanished (22 taxa). In the previous catalogue, they were mostly classified as naturalized, often post-invasive (Pyšek et al. 2002); the change in classification of these taxa resulted from the focus on the current state adopted in the present treatment and the fact that they no longer occur in populations that can be considered self-sustaining. The group includes some red-listed archaeophytes (e.g. *Agrostemma githago*, *Atriplex rosea*, *Heliotropium europaeum*, *Lolium remotum* and *Scandix pecten-veneris*; Holub & Procházka 2000), but also neophytes (e.g. *Cnidium silaifolium* and *Xanthium spinosum*), and refers to the invasion failure in the sense of Blackburn et al. (2011).

**Past link to cultivation**

**Group 3. Establishment and failure.** A group of 17 taxa that are either archaeophytes or neophytes introduced long ago, mostly in the 19th century, were surviving due to weak but continued propagule pressure from cultivated populations in the past but never formed self-sustaining population in the wild. Since the planting has ceased or its intensity strongly decreased, they are currently declining or have already vanished (13 taxa). Examples include *Camelina sativa*, *Chenopodium foliosum*, *Dracocephalum moldavica*, *Madia sativa*, *Pyrus nivalis*, *Stachys affinis* or *Trigonella foenum-graecum*.

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**Fig. 3.** – Population groups (PG) of alien taxa in the Czech flora (see text for details and Table 1 for overview) mapped onto the unified framework for biological invasions (Blackburn et al. 2011; the background figure reprinted with permission from Elsevier Limited). Population groups corresponding to casual , naturalized but not invasive , and invasive taxa are distinguished by different colours. Number of taxa and percentages of the total of 1454 are indicated for each stage. Note that the groups do not match precisely the casual–naturalized–invasive areas at the top of the scheme due to distinguishing taxa that correspond to invasion boom and bust (taxa that spread in the past, formed metapopulations but their spread ceased, therefore are at present considered naturalized rather than invasive; PG13+15+17).
(a3) **Ongoing link to cultivation**

**Group 4 & 5. Introduction and failure.** An escape from cultivation analogous to Group 1. Group 4 includes 501 casual taxa, mostly neophytes (458), that rely on continued input of propagules from planted populations. Usually they are planted as garden ornamentals and the link between planted populations and those in the wild is very close. In terms of abundance, these taxa are at best scattered (339 are rare, 109 reported from a single site) and 56 are vanished. Examples include *Convolvulus tricolor*, *Dahlia pinnata*, *Dasiphora fruticosa* and *Ficus carica*. Some woody plants that escaped from cultivation have close link with planted populations, but have not formed (yet) long-sustaining populations due to long generation time (e.g. *Celtis occidentalis*, *Crataegus persimilis* and *Paulownia tomentosa*) or limited ability to establish permanently (e.g. *Abies grandis* and *Platanus × hispanica*) are included in this group. Some taxa previously classified as naturalized by Pyšek et al. (2002) were reassigned to this group (e.g. *Allium tuberosum*, *Helleborus viridis*, *Othocallis siberica*, *Polygonatum latifolium* and *Sedum rupestre subsp. erectum*), including some shrubs surviving in single or a few locations (e.g. *Alnus rugosa*, *Ribes odoratum* and *Rubus canadensis*).

Group 5 is defined based on the same principles, the difference being current rather massive propagule pressure from large-scale planting for agricultural or horticultural purposes. It includes 28 taxa, with archaeophytes prevailing (21) but neophytes also represented, and examples include *Allium cepa*, *Anethum graveolens*, *Helianthus annuus*, *Triticum aestivum* or *Zea mays*. There are 18 anecophytes in this group.

(b) **Self-sustaining isolated populations**

(b1) **No link to cultivation**

**Group 6. Establishment and failure.** This group includes 54 archaeophytes that were introduced independently of cultivation, survived in the landscape for centuries or even millennia and although their populations are declining, they still survive in the wild as rare or scattered. The majority of them occur in warm regions and it is assumed that many of them were invasive at some stage in their invasion history (classified as naturalized post-invasive in Pyšek et al. 2002), often as weeds of arable land. Examples include *Ajuga chamaepitys subsp. chamaepitys*, *Anagallis foemina*, *Bifora radians* and *Ranunculus arvensis*. A subset in this group are taxa confined to habitats associated with breeding domestic animals in villages, e.g. *Chenopodium vulvaria*, *Lepidium coronopus*, *Marrubium peregrinum* and *Sclerochloa dura*.

**Group 7. Establishment and no trend.** The group consists of 40 taxa, most of them archaeophytes (21) but also old neophytes are represented (19), most of them introduced in the 19th century. The taxa from this group occur mostly as scattered or rare but without a significant trend for decline or spread. Examples include: *Brachypodium rupestre*, *Genista sagittalis*, *Crepis capillaris*, *Geranium molle*, *Papaver dubium*, *Pastinaca sativa subsp. urens* and *Potentilla intermedia*.

**Group 8. Starting spread.** A group comprising almost exclusively neophytes (40 of 43 in total), mostly introduced in the 20th century, that have formed self-sustaining populations and exhibited signs of starting spread in the last decades. The majority of them were classified as naturalized in the previous catalogue (Pyšek et al. 2002), but there are also 11 taxa that were in the casual stage at the beginning of the 2000s and their dynamics in the last decade justifies reassessment, e.g. *Abutilon theophrasti* and *Senecio inaequidens*. The
group includes also taxa that formed a small but abundant and persisting population that is currently prevented from further spread by the barrier of unsuitable habitats (*Corispermum pallasii*) or those that were introduced fairly recently and had not time yet to fully manifest their invasion potential (*Agrostis scabra*, *Dittrichia graveolens* and *Panicum miliaceum* subsp. *agricola*).

**b2** Past link to cultivation.

**Group 9. Establishment and no trend.** An escape from cultivation analogous to Group 7. This group includes 36 taxa, mostly neophytes (27), that form stabilized self-sustaining populations in the wild as a result of past planting, ranging from rare to common in abundance (e.g. *Calystegia pulchra*, *Hesperis matronalis* subsp. *matronalis*, *Saxifraga hostii* subsp. *hostii* and *Viola suavis*), but also archaeophytes with the same characteristics (*Glycyrrhiza glabra*, *Lilium bulbiferum* and *Myrrhis odorata*).

**Group 10. Starting spread.** This group includes 11 taxa, nine of them being naturalized neophytes that exhibit signs of starting spread and are likely to become invasive in the future, e.g. *Dipsacus strigosus* and *Duchesnea indica*. Compared to previous catalogue (Pyšek et al. 2002), *Azolla filiculoides* and *Bromus carinatus* that were assessed as casual, appear in this category. The group also includes two archaeophytes, *Bryonia dioica* and *Galega officinalis*.

**b3** Ongoing link to cultivation

**Group 11. Establishment and no trend.** A group of 65 taxa with early introduced neophytes prevailing (57 taxa, for the majority of them the first record is available from the 19th century), that occur as rare or scattered but have formed self-sustaining populations with ongoing support of propagule pressure from cultivated populations. Examples include *Alcea rosea*, *Lychnis coronaria* and *Matteuccia struthiopteris*. Compared to previous classification (Pyšek et al. 2002), 25 taxa considered as casual then are now considered to form self-sustaining populations, e.g. *Arabis procurrens*, *Eranthis hyemalis* and *Erysimum cheiri*. Populations of some taxa are likely to start spread in the future, being currently still constrained by a short residence time (e.g. *Elaeagnus commutata*).

**Group 12. Starting spread.** This group includes 31 taxa, all but one neophytes, that are still more or less widely planted and exhibit the signs of beginning spread, e.g. *Colutea arborescens*, *Fallopia auberti*, *Hordeum jubatum* and *Pinus nigra*. Based on the marked dynamics in the last decade, some of them were reclassified from the casual category in Pyšek et al. (2002) to naturalized, e.g. *Buddleja davidii* (first reported to escape from cultivation in 2000), *Aesculus hippocastanum*, *Symphyotrichum laeve* or *Sagittaria latifolia*. The group also includes several taxa formerly classified as invasive for which this classification is not (yet) justified using the conservative approach adopted here: they are *Amorpha fruticosa*, *Cytisus scoparius* subsp. *scoparius*, *Galeobdolon argentatum*, *Mahonia aquifolium*, *Physocarpus opulifolius*, *Rhus typhina* or *Sedum hispanicum*.

**c** Invasive metapopulations  
**c1** No link to cultivation

**Group 13. Establishment and no trend.** A group of 100 unintentionally introduced taxa with occurrence stabilized during centuries or millennia of presence in the target region, consisting mostly of archaeophytes (87 taxa). The examples include many common weeds
of agricultural land and ruderal taxa such as *Anagallis arvensis*, *Anthemis arvensis*, *Chenopodium strictum*, *Convolvulus arvensis*, *Euphorbia peplus*, *Lamium purpureum*, *Lapsana communis* subsp. *communis*, *Malva neglecta* and *Thlaspi arvense*. Majority of taxa (68) were assumed to be post-invasive in Pyšek et al. (2002). Sixteen species previously classified as invasive were reassigned into this naturalized category, e.g. *Apera spica-venti*, *Atriplex oblongifolia*, *Bryonia alba*, *Epilobium adenocaulon*, *Matricaria discoidea*, *Rumex thyrsiflorus*, *Tripleurospermum inodorum* and *Veronica persica*.

**Group 14. Spread.** This group includes 28 taxa that became invasive following unintentional introduction. Most of them are neophytes (20), e.g. *Amaranthus powellii*, *Ambrosia artemisiifolia*, *Bidens frondosus*, *Cynara canadensis*, *Cuscuta campestris*, *Rumex alpinus*, but invasive archaeophytes are also represented, e.g. *Atriplex sagittata*, *Cirsium arvense*, *Echinochloa crus-galli* and *Portulaca oleracea* subsp. *oleracea*. Apparently, annual weeds prevail with some exceptions such as *Bunias orientalis*, whereas both other invasive groups (16 and 18) consist mainly of robust perennials and woody taxa, the differences reflecting life histories associated with unintentional vs deliberate pathways of introduction (Pyšek et al. 2011b).

(c2) Past link to cultivation

**Group 15. Establishment and no trend.** Group of eight taxa, both archaeophytes (e.g. *Cymbalaria muralis* and *Spergula arvensis* subsp. *sativa*) and neophytes (e.g. *Acorus calamus* and *Elodea canadensis*), with the same features as Group 13 but supported in their naturalization by past cultivation, and no longer spreading. *Elodea canadensis*, *Mimulus guttatus*, *Tanacetum vulgare* and *Veronica filiformis* have been reclassified from invasive status (Pyšek et al. 2002) to naturalized.

**Group 16. Spread.** Nine taxa that still spread and the naturalization and invasion of which has been supported by planting that was most intensive in the past; they are all early introduced neophytes classified as invasive already in the previous catalogue (Pyšek et al. 2002): *Ailanthus altissima*, *Angelica archangelica* subsp. *archangelica*, *Echinops sphaerocephalus*, *Heracleum mantegazzianum*, *Impatiens glandulifera*, *I. parviflora*, *Lycium barbarum* and *Telekia speciosa*. The only exception is *Asclepias syriaca*, previously classified as naturalized; this species started to spread in the last decade, especially in southern Moravia.

(c3) Ongoing link to cultivation

**Group 17. Establishment and no trend.** A group of 19 taxa, consisting of 12 archaeophytes and 7 neophytes that are still commonly planted at present and form stabilized metapopulations in the wild. Examples include *Armoracia rusticana*, *Lolium multiflorum*, *Prunus cerasus* and *Trifolium hybridum*. Twelve taxa were classified as post-invasive by Pyšek et al. (2002) and four considered as invasive in this source were reassessed (*Digitalis purpurea*, *Melilotus albus*, *M. officinalis* and *Viola odorata*) and included in this group of naturalized taxa.

**Group 18. Spread.** A group of 24 invasive taxa that are currently spreading were supported by planting throughout their invasion history, including the present time. There are only two archaeophytes, *Arrhenatherum elatius* and *Prunus cerasifera*, while the vast majority of species in this group are neophytes that started to appear in the wild in the 19th century. The examples include many major plant invaders in the Czech Republic such as
Acer negundo, Helianthus tuberosus, Lupinus polyphyllus, Pinus strobus, Prunus serotina, Quercus rubra, Reynoutria ×bohemica, R. japonica var. japonica, Robinia pseudacacia, Solidago canadensis and S. gigantea. All taxa in this group but Prunus cerasifera were classified as invasive already in Pyšek et al. (2002). Although taxa confined to eutrophic ruderal habitats generally prevail in this group, those preferring nutrient-poor soils (such as Pinus strobus, Prunus serotina, and Quercus rubra) are also present.

Taxonomic composition

Alien taxa in the Czech flora are representatives of 586 genera and 107 families (Appendix 2). The genera richest in taxa (including hybrids and anecophytes) among all aliens are Amaranthus (24 taxa), Oenothera (23) and Trifolium (19) but there are marked differences between neophytes and archaeophytes in this respect: Oenothera, Amaranthus, Trifolium, Rumex, Solanum, Rubus and Centaurea are most represented genera among neophytes, whereas Vicia, Prunus, Veronica, Atriplex, Bromus, Viola and Chenopodium among archaeophytes (Table 4).

Overall, neophytes belong to 508 and archaeophytes to 184 genera; exclusively ‘archaeophytic genera’ (with only archaeophytes among their alien taxa) that include at least three alien representatives are Arctium (7 taxa), Spargula (4), Anthriscus, Marrubium, Myosotis, Polycnemum, Pyrus, Sonchus and Valerianella (3).

Families most represented in alien flora (Table 5) are Asteraceae (198 taxa; 13.6% of the alien flora), Poaceae (152; 10.5%) and Brassicaceae (101; 6.3%); apart from minor changes in the numbers of taxa resulting from the above described additions and deletions, the pattern of richness at the level of most represented families is the same as reported in detail in Pyšek et al. (2002). Some major changes in the richness of families in the current treatment, compared to Pyšek et al. (2012; e.g. Amaranthaceae 76 vs 25 taxa, Scrophulariaceae 5 vs 39), are attributed to the different classification system used here (Stevens 2001 onwards, The Angiosperm Phylogeny Group 2009). All but one (Linaceae) of the total number of 107 families included contain at least one neophyte representative, while archaeophytes originate from only 42 families. The families richest in neophytes are Asteraceae, Poaceae, Rosaceae, Fabaceae and Brassicaceae (Table 5), which together contain 485 taxa and account for 43.9% of all neophytes. Asteraceae, Poaceae and Brassicaceae also rank high among archaeophytes, but there are also other families that are rich in archaeophytes (e.g. Apiaceae, Caryophyllaceae, Plantaginaceae and Boraginaceae; Table 5).

Temporal trends and pathways of introduction

The data on the first record in the studied region, known for 771 neophytes, allow to reconstruct the increase in the number of taxa introduced into the Czech Republic over the last three centuries, although it is clear that the reliability of data on residence times decreases towards the past (Lambdon et al. 2008). The numbers of new taxa recorded in particular years reflect peaks associated with specific events such as the increased interest in plants of human-made habitats in the 1970s, linked to the establishment of a working group at the Institute of Botany (Hejný et al. 1973, Pyšek 2001, Pyšek et al. 2003, 2011b), or the publication of the first catalogue of Czech alien plants (Pyšek et al. 2002). However, when the cumulative number of the first species records is plotted against time, the trend suggests a rather steady increase of four alien arrivals per year since the beginning of the 19th century...
without any distinct decelerating trend and a projected total number of 1264 taxa in the year 2050. Fifty per cent of the present known taxa were recorded up to 1935, 60% up to 1957, 70% up to 1963, 80% up to 1973, and 90% up to 1997 (Fig. 4). This indicates that the number of alien taxa recorded in the Czech Republic will be increasing at a similar rate in the near future, corresponding to a trend reported for Europe (Hulme et al. 2009) and creating an invasion debt (Essl et al. 2011).

As to the pathways of introduction into the country, deliberate introduction was involved in 747 of the 1454 taxa (51.4%). Most deliberate introductions resulted from ornamental or horticultural plantings (see Pyšek et al. 2002 for detailed analyses of planting purposes). The remaining 48.6% of taxa are assumed to have arrived by unintentional pathways, i.e. mostly as contaminants of commodities or stowaways (Hulme et al. 2008, Pyšek et al. 2011b). The ratio of deliberate and unintentional introduction is reversed in archaeophytes and neophytes, with 30.7% of the total number of taxa deliberately introduced among the former and 57.9% among the latter.

Table 4. – Genera with the highest diversity of alien taxa in the Czech flora, cross-tabulated according to immigration time and invasion status. The 23 genera represented by at least 10 alien taxa are shown. Other taxon-rich genera include Avena, Cirsium, Hordeum, Malva, Papaver, Setaria, Silene, Sisymbrium, Symphyotrichum (8 alien taxa), Brassica, Camellina and Fumaria (7 alien taxa). Hybrids are included. Cas – casual; natur – naturalized non-invasive; inv – invasive.

<table>
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<th>Genus</th>
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<th>Neophytes</th>
<th>Total</th>
</tr>
</thead>
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Table 5.—Families with the highest diversity of alien taxa in the Czech flora, cross-tabulated according to immigration time and invasion status. The 29 families represented by at least 10 alien taxa are shown. Hybrids are included. Cas – casual; natur – naturalized but non-invasive; inv – invasive. The classification of families follows that of Angiosperm Phylogeny Group: APG III (Stevens 2001 onwards, Angiosperm Phylogeny Group 2009).

<table>
<thead>
<tr>
<th>Family</th>
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<th>Neophytes</th>
<th>Total</th>
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**Life histories and regions of origin**

Among all aliens, 43.3% are annuals, 33.1% perennials, 10.8% biennials, 8.5% shrubs or semishrubs, and 4.3% trees. Archaeophytes and neophytes demonstrate a highly significant difference in the distribution of life histories: the former are more often annuals (56.4% vs 38.8% among neophytes) or biennials (17.0% vs 8.6%) and less often perennials (18.2% vs 38.3%) or shrubs and trees (8.5% vs 14.3%; Fig. 5).

The main donors of alien plants to the Czech Republic are the Mediterranean region (34.6%), other parts of Europe (19.4%), other parts of Asia (13.1%) and North America (12.6%). The contribution of other regions (Central America, South America, Africa, Australia) does not exceed 4%. The region of origin could not be assigned for 199 taxa, a group consisting of 105 anecophytes and 94 taxa of hybrid origin (Fig. 6). The data on origins confirm the well-known difference between archaeophytes and neophytes in terms
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Fig. 4. Temporal trends in the alien flora of the Czech Republic in the last 200 years based on neophytes with known year of the first report (n = 771). Also shown is extrapolated trend for the total number of taxa (n = 1104), and numbers of taxa reported in particular years (right axis).

Fig. 5. – Representation of life histories among alien taxa in the Czech Republic. Taxa with multiple life histories were considered in each category so the sum of the numbers of taxa (shown on top of the bars) does not match the total numbers of archaeophytes and neophytes. Overall, the observed counts of alien taxa highly significantly ($\chi^2 = 94.25; df = 4; P < 0.0001$) differ from counts expected by chance (values in parentheses). Statistically significant deviations of individual counts from counts that can be expected by chance are expressed by the number of asterisks (*** P < 0.001; ** P < 0.01; * P < 0.05) and marginal significance by a dot (. P < 0.1); numbers in parentheses not followed by any symbol do not differ from randomly expected values. Semishrubs are included within shrubs. Excluded from these statistics are 4 ferns (all neophytes), 11 aquatic species (all neophytes) and 11 parasitic species (3 archaeophytes, 8 neophytes).
of source regions (e.g. Pyšek et al. 2002, 2004b, 2005, Chytrý et al. 2005, 2008a, b): more than a half (52.7%) of archaeophytes originate from the Mediterranean region (the figure increases to 64.5% if anecophytes and hybrids are excluded), which is, however, also the most frequent donor of neophytes (28.7%). The contribution of other parts of Europe and Asia to the total number of taxa is slightly higher for neophytes than for archaeophytes, 19.9% vs 17.8% and 14.2% vs 10.1%, respectively (Fig. 7).

Since archaeophytes, by definition, have not arrived from overseas, it is plausible to compare their regions of origins with those of neophytes if Americas and Australia are excluded. The difference between archaeophytes and neophytes in such a comparison is still statistically highly significant ($\chi^2 = 45.057; df = 3; P < 0.0001$). Highly significantly ($P < 0.001$) more archaeophytes originated in the Mediterranean region (231 vs 180.5 expected counts), but highly significantly less ($P < 0.01$) in the other parts of Asia (44 vs 67.9), significantly ($P < 0.05$) less in the other parts of Europe (78 vs 100.1) and marginally significantly less ($P < 0.1$) in Africa (5 vs 9.6). Conversely, neophytes originated in the Mediterranean region were significantly less represented (385 vs 436.5) and those from the other parts of Asia marginally significantly more represented (190 vs 164.1).
Regional abundance, habitats and cover in plant communities

Archaeophytes are generally more abundant in the field, which reflects that they were provided with more time in the target region (Pyšek et al. 2002, 2004b, 2011a). Of the total number of archaeophytes, 22.0% are considered common (highly significantly more than expected by chance), 2.9% locally abundant and 28.5% scattered (highly significantly more than expected by chance). This pattern strikingly contrasts with that found for neophytes. Only 2.9% of neophytes (35 taxa) are classified as common (highly significantly less than expected by chance) and 3.0% locally abundant, 8.1% scattered (highly significantly less than expected) while as many as 86.0% occur in low-abundance categories (rare, single locality or vanished; with the last two categories occurring highly significantly or significantly, respectively, more often than expected by chance); the corresponding figure for archaeophytes being 46.6%, with these categories significantly or highly significantly underrepresented. Two hundred and twelve neophytes (17.7%) are only known from a single locality (compared to only five archaeophyte hybrids; Appendix 2) and 250 (22.6%) are labelled as vanished (compared to only 27 archaeophytes, i.e. 7.7%) (Fig. 8).

The contrasting patterns in the occurrence of both immigration status groups, archaeophytes and neophytes, translate into those of the breadth of their habitat niches, expressed as the number of habitats of the total of 88, occupied by 497 taxa that could be classified according to their habitat affinities (Sádlo et al. 2007). Archaeophytes occupy on average more habitats (9.5±9.0, mean±S.D., n = 244) than neophytes (6.4±6.1, n = 253), and 31.6% of them occur in more than 10 habitats (compared to only 17.8% of

Fig. 7. – Distribution of archaeophytes and neophytes in the Czech Republic according to their origin. Taxa originating from multiple regions as designated here are included in each region. See text for the results of statistical analysis.
neophytes; Fig. 9). Ten archaeophytes and only three neophytes (Conyza canadensis, Epilobium adenocaulon and Impatiens parviflora) grow in a wide range of habitats exceeding 30 (see Sádlo et al. 2007: their Table 2). The species with the broadest habitat niche of all alien taxa in the Czech Republic is an archaeophyte, Arrhenatherum elatius, occurring in 62 of 88 habitats (see Appendix 1 for comments on its classification).

The covers that alien taxa reach in plant communities in the Czech Republic yield a completely opposite picture of neophyte vs archaeophyte comparison (Fig. 10). Neophytes are shifted towards high-cover categories, reaching on average 8.5% cover (n = 48), markedly more than archaeophytes (4.7%, n = 131). The first five taxa with highest average covers are all neophytes: Acorus calamus 39% (recorded in n = 293 vegetation plots), Elodea canadensis 35% (n = 412), Helianthus tuberosus 26% (n = 62), Heracleum mantegazzianum 26% (n = 27) and Reynoutria japonica var. japonica lumped with R. ×bohemica 26% (n = 51). Other neophytes with a high cover are Impatiens glandulifera (18%, n = 302), Solidago gigantea (17%, n = 99), Echinocystis lobata (14%, n = 33) and Pinus nigra (13%, n = 33).
Fig. 9. – Frequency distribution of the numbers of habitats (n = 88) in which alien taxa are recorded, shown separately for archaeophytes (n = 244) and neophytes (n = 253).

Fig. 10. – Frequency distribution of covers of alien taxa in plant communities in the Czech Republic. Only taxa for which data from at least 25 plots are available were included. Numbers of taxa in each cover class are shown on top of the bars.
Although this comparison must be taken with caution because the vegetation plots were sampled in a subjective, preferential way, average plot sizes for individual taxa differ and there is also great variation in the number of plots from which the data are derived, the differences between the two groups of aliens are robust enough to indicate that neophytes are on average more successful in colonizing plant communities and often forming monodominant stands (see also Chytrý et al. 2008a).

Impact

A thorough assessment of impacts of plant invasions in the Czech Republic is still missing which reflects the fact that studies summarizing information on impacts across alien floras of large regions are still rare despite intensive research in the last few years (Parker 1999, Gaertner et al. 2009, Pyšek & Richardson 2010, Vilà et al. 2010, 2011, Winter et al. 2009, Pyšek et al. 2012). Based on data on impacts of alien plants in Europe summarized by the DAISIE project (DAISIE 2009, www.europe-aliens.org), there are 133 taxa on the list of Czech alien plants that were documented in the literature to exert ecological impacts and/or economic impacts in some parts of Europe (Appendix 2), some of them also in the Czech Republic (Hejda et al. 2009). These data make it possible to highlight taxa that already impose ecological impacts but also those that can become threat in the future.

The group of taxa with documented ecological impacts covers 33 taxa that are classified as invasive in the present study, and includes most of the major invaders in the Czech Republic, some of them threatening seminatural habitats (e.g. Acer negundo, Ailanthus altissima, Helianthus tuberosus, Heracleum mantegazzianum, Impatiens glandulifera, Impatiens parviflora, Lupinus polyphyllus, Lycium barbarum, Pinus strobus, Prunus serotina, Reynoutria japonica var. japonica, R. sachalinensis, R.×bohemica, Robinia pseudoacacia, Rudbeckia laciniata, Solidago canadensis and S. gigantea) but also noxious weeds of arable land (e.g. Amaranthus retroflexus and Galinsoga parviflora) or species affecting human health (Ambrosia artemisiifolia). Besides these taxa, already exerting impacts in the Czech Republic, the 113 taxa with ecological impacts in Europe include 45 that we currently classify as naturalized; some of them belong to population groups that exhibit symptoms of starting spread and their impact in the near future is likely (e.g. Abutilon theophrasti, Lepidium virginicum and Senecio inaequidens). Finally, for 35 taxa that occur as casual in the Czech Republic ecological or economic impact is documented from elsewhere in Europe; this group includes some noxious invaders (e.g. Elodea nuttalii, Rosa rugosa and Solidago graminifolia) that should be monitored to enable early action should their population dynamics change (Appendix 2).

Notes on the classification of taxa

The present update of the 10 years old data yielded a number of changes to the taxa listed, and their invasion and residence time statuses. These changes are due to several reasons. First, they reflect the real changes in species’ behaviour and their invasion dynamics over the last decade. Second, the interest in and knowledge of alien plants has improved considerably as a result of intensive research in biological invasions in the Czech Republic during this period. Third, the more conservative approach towards what should be considered native or alien also brought about changes in the species list, and finally, introducing the
population-based approach to the classification of taxa adopted here (Blackburn et al. 2011) resulted in shifts in invasion status.

The main change in approach relative to the previously used scheme concerns a strict focus on the current state of a taxon’s populations in a region. This allowed us to take into account and quantify categories that refer to unsuccessful invasions – the ‘invasion failure’ and ‘boom and bust’ phenomena as defined by Blackburn et al. (2011). This is reflected namely in classifying taxa that formed self-sustained populations in the past, some assumed to have been invasive (and labelled post-invasive in Pyšek et al. 2002), as casual, suggesting the reversed trajectory along the INIC (Fig. 2). Although they would not be classified as casuals, should the criterion of relying on repeated introduction of propagules, which is part of the traditionally accepted definition, be strictly followed (Richardson et al. 2000, 2011), we believe that the criterion of population self-sustainability is a more important one, reflecting closely the population dynamics in both directions along the INIC. This approach is further supported by the fact that many of these taxa are red-listed or missing for a long time, which strongly argues against self-sustainability of their populations. This group includes also many archaeophytes that have never been planted indicating that their occasional occurrence is due to long-term survival in and occasional germination from seed banks.

Consequently, the number of invasive taxa is substantially smaller than in the previous catalogue (50 neophytes and 11 archaeophytes in the present study compared to 69 neophytes and 21 archaeophytes, respectively, in Pyšek et al. 2002). A decrease this dramatic is due to the newly adopted conservative approach; unlike in the previous account, the emphasis here was on ongoing spread as a major criterion. The lower numbers do not mean that the problems with invasive plants in the Czech Republic are diminishing; rather the opposite is true as indicated by species that started to spread recently. In conclusion, we believe that the more rigorous approach to separating invasive species from naturalized makes the current assessment of species status more comparable with other parts of the world, especially those that experience serious problems with invasions, and forms a sounder basis for managing plant invasions at the national scale.

See http://www.preslia.cz for Electronic Appendices 1,2

Acknowledgments

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Souhrn


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Vymyslický T. & Grulich V. (2004): *Chorispora tenella* a *Corispermum canescens* na jižní Moravě [*Chorispora* *tenella* and *Corispermum* *canescens* found in South Moravia]. – Masarykova univerzita, Brno.


Received 29 May 2012
Revision received 8 June 2012
Accepted 11 June 2012
Appendix 1. – Comments on taxa that represent changes against the previous Catalogue of alien plants of the Czech Republic (Pyšek et al. 2002).

Changes of names, difficult cases and corrections of earlier misidentifications

Compared to the previous version of the catalogue (Pyšek et al. 2002), 124 names were changed due to nomenclatural reasons or development in taxonomic opinion (Electronic Appendix 1). Additional seven taxa are listed under a different name due to the reidentification; their names refer to the same taxa which were erroneously determined in 2002 or their taxonomic classification has changed. These taxa are commented below and represent additions to the alien flora of the country.

Azolla filiculoides was listed as A. caroliniana in Pyšek et al. (2002), based on treatment in the Flora of the Czech Republic (Krása in Hejní & Slavík 1988). The taxonomy of the New World Azolla has been controversial for a long time. The number of distinguished species varied and different characters were used for their identification. However, Evrard & Van Hove (2004) in their recent thorough investigation based on morphological, molecular and physiological data concluded that only two species can be distinguished taxonomically in America. They revealed that the type specimen of A. caroliniana belongs to the species described earlier as A. filiculoides, and the fern usually identified as A. caroliniana by many authors should be correctly named A. cristata. Although both species were recorded as introduced in Europe, only A. filiculoides is widespread, whereas A. cristata was apparently documented only from the Netherlands. Plants recently collected in the Czech Republic are identical with A. filiculoides (coll. and det. Z. Kaplan, PRA, rev. C. Van Hove). The other species, A. cristata (A. caroliniana auct.), has apparently never occurred in the country as introduced or escaped.

Corispermum pallasi was listed in Pyšek et al. (2002) as C. leptopterum. However, recent taxonomic studies revealed that the European plants are conspecific with the Siberian ones, described much earlier as C. pallasi (Mosyakin 2003). Vymyslický & Grulich (2004), reporting on their find of Corispermum from Ivančice, Brno, suggested that southern Moravian plants correspond to C. canescens, which is native to Hungary. However, based on a careful re-examination of specimens from BRNU and PR (J. Danihelka), we believe that all Corispermum specimens so far collected in the Czech Republic, with the only exception of C. declinatum (see below), most likely belong to C. pallasi. The earliest documented record of this species is from 1933 (ex herb. F. Hrobař, PR). At present, C. pallasi occurs in two populations consisting of thousands of plants in sand pits near Bzenec, southern Moravia, from where it is spread with traded sand to other places.

Eriochloa punctata. Three Eriochloa species were reported in the literature from the Czech Republic: E. ramosa from a wool-processing factory Mosilana in Brno (Dvořák & Kühn 1966, Grüll 1979) and E. punctata from railway station in Brno (Grüll 1979); these two species are given in the Flora of Dostál (1989), who in addition lists E. procera, all as casual wool aliens introduced to Brno. Actually, the names E. procera and E. ramosa refer to the same taxon, with the former accepted as its correct name (Zuloaga & Morrone 2003, Shouliang & Phillips 2006). The plant reported as E. ramosa by Dvořák & Kühn (1966), collected by F. Kühn in 1960, was deposited in BRNU in 1972 under the name E. punctata; obviously, J. Dvořák re-determined the plant before depositing it in the herbarium. Comparison of the specimen collected by F. Kühn in 1960 and another specimen collected by F. Grüll in 1965 (reported by Grüll 1979) has shown that both of them very likely represent the same species, most probably E. punctata, as already suggested by J. Dvořák (rev. J. Danihelka). Consequently, the species listed as E. procera in Pyšek et al. (2002) is in fact E. punctata, the same as found by Grüll (1979).

Gilka achilleifolia. Another species of the genus, G. multicaulis, is listed in Pyšek et al. (2002), based on a note in the Flora of the Czech Republic (Krása in Slavík 2000) that it is planted and rarely escapes from cultivation, without further details. In 2005, two flowering plants of G. achilleifolia were reported growing in the Stářkovský les forest near Lanžhot, southern Moravia, on a forest clearing along a road, together with Legousia pentagonia. They were probably introduced to the site with forestry vehicles (Řehořek & Lososová in Hadinec & Lustyk 2009). Since G. multicaulis is sometimes classified as G. achilleifolia var. multicaulis, we included only G. achilleifolia as it is possible that the above reports refer to this taxon.

Hieracium heldreichii agg. Listed in Pyšek et al. (2002) as H. pannosum, a cultivation relict. In the 1930s it was introduced at Kunětická hora hill near Pardubice (eastern Bohemia) and is still persisting there. The species, originally determined as H. pannosum by J. Holub, has been now re-identified by Z. Szela as Hieracium sp. ex H. heldreichii agg. The species is native to the Balkan Peninsula.

Rodgersia pinnata. The species reported in Pyšek et al. (2002: Fig. 1b) as R. aesculifolia was misidentified. The mistake was corrected by Král et al. (2004c).

Spiraea hypericifolia subsp. obovata was listed as S. crenata in Pyšek et al. (2002), based on the escape from cultivation reported from ruins of the Škalka castle near Vlastislav, northern Bohemia, at the end of the 19th cen-
tury (Koblížek in Hejný & Slavík 1992). This report was based on erroneous determination of a herbarium specimen that belongs to *S. hypericifolia* subsp. *obovata* (Businský & Businská 2002).

**New taxa: additions to the alien flora of the Czech Republic**

The following 151 taxa, not listed in Pyšek et al. (2002), represent additions to the alien flora of the Czech Republic:

- *Abies concolor*, *A. grandis* and *A. nordmanniana*. Natural regeneration from seed produced by planted trees occurs in the Průhonice Park near Prague (J. Burda, pers. comm.).
- *Acanthus hungaricus*. A rarely planted species in the Czech Republic, first recorded as escaped from cultivation in Praha (Prague)-Lipence in 1999. The population of ca 150 flowering plants, reproducing by seed and surviving winter, probably resulted from planting along the wall of a baroque farmstead in the early 1990s (Hadinec in Hadinec & Lustyk 2009).
- *Acer tataricum* was part of the Czech flora in the Subatlantic period but became extinct (Opravil 1967). Its modern presence is due to escapes from cultivation and subsequent naturalization, with the first record of planting in 1835 (Koblížek in Slavík 1997a). In 2004, it was recorded regenerating in the Hevlínské jezero Nature Reserve, distr. Znojmo, southern Moravia (Čáp & Koblížek in Hadinec et al. 2005 as var. *torminaloides*). Self-sown plants established from seed were further observed in the vicinity of planted individuals in numerous locations in Prague (Suchdol, Černý Most and Libeň; recorded by J. Sádlo in 2011–2012).
- *Actinidia deliciosa* was first recorded in the wild at the channelled stream of Botič in Prague in 2008, forming a population of seven sterile plants, originated from seed of the kiwi fruit (Hadinec et al. in Hadinec & Lustyk 2008). The first report in Europe of its occurrence outside cultivation is from Germany in 1998 (Kasperek 2003), followed by records in other countries and natural habitats. Seeds germinate well, plants spread vegetatively and survive mild winters. Populations in suitable climatic conditions can be therefore considered as likely candidates for naturalization and spread (Hadinec et al. in Hadinec & Lustyk 2008).
- *Ageratina altissima* occasionally escapes from cultivation, with so far the single documented record from the vicinity of the Ostravice railway station in northeastern Moravia in 1979 (Slavík in Slavík & Štěpánková 2004).
- *Allium christophii*. A commonly planted species, found in a scrub near Kostomlát, central Bohemia, in 1994. Occasional escapes from gardens can be expected also in other places, but since the bulbs are consumed by rodents, its naturalization is unlikely (Krahulec & Lepší in Hadinec & Lustyk 2009).
- *Allium roseum*. In 2005 a population of 18 plants was recorded in Hojná Voda, southern Bohemia, at a site that is probably a long abandoned garden, and it was still present there in 2009; further spread cannot be excluded as the species is a prolific bulbil producer (Krahulec & Lepší in Hadinec & Lustyk 2009).
- *Allium stipitatum*. A frequently planted species, found once escaped from cultivation along a road in Hradčany, central Bohemia, in 2008. Occasional escapes from gardens can be expected due to frequent planting but naturalization is unlikely because bulbs are consumed by rodents (Krahulec & Lepší in Hadinec & Lustyk 2009).
- *Allium zebdanense*. First documented from abandoned garden allotments in Praha-Střížkov in 2006, but the species is known to grow spontaneously for several decades in the Botanical Garden of Charles University in Prague. It has not been reported escaped from cultivation from other central European countries as yet, but further records from the wild are likely to appear in the future because plants produce a number of small bulbils providing the species with potential to spread (Krahulec & Marek in Hadinec & Lustyk 2006).
- *Amelanchier alnifolia* was first recorded outside cultivation in Český Krumlov, southern Bohemia, in 2008 (Lepší & Lepší 2008), but it was uncertain if the plants were escapees from cultivation or remnants from planting. Since then it has been repeatedly observed as escaping from cultivation (M. Lepší, pers. comm.).
- *Amelanchier spicata*. First documented from the wild by a herbarium specimen collected near Havlíčkův Brod in 1880, the species was recently reported from 32 localities scattered over the country, growing naturalized in scrub, oak and pine forests, their margins and in river valleys. A recent review revealed that it is the most frequently planted and escaping species of the genus in the Czech Republic (Lepší & Lepší 2008).
- *Ammobium alatum* is planted as an ornamental plant and rarely escapes from cultivation (Slavíková in Slavík & Štěpánková 2004). Outside cultivation it is reported from several sites in the Železné hory Mts, with the first record from 1942 (Hadač et al. 1994), and from a ruderal site in Bruntál, northern Moravia (Hradilek et al. 1999).
- *Amsinckia lycopsoides* was found once growing in a lawn in Brno-Bohunice in 2000, probably introduced with soil or as a seed admixture. The population was destroyed by planting of shrubs in the following year (Rotreklová & Řehořek in Hadinec & Lustyk 2009). It was also growing on rocks adjacent to a private garden in Vimperk, southwestern Bohemia, following an unintentional introduction. It survived there for several years in the 1990s (F. Krahulec, pers. obs.).
Anthemis cotula × Cota tinctoria (syn. Anthemis ×bollei). This hybrid was found once in Břeclav-Boštorná, southern Moravia (1994, BRNU; Dvořáková in Slavík & Štěpánková 2004).

Anthemis cretica subsp. columnae. Status of this taxon in the Czech botanical literature is unclear. It was reported from three localities since 1871, last observed in the 1920s (Dvořáková in Slavík & Štěpánková 2004). Given the scattered distribution in the mountains of southwestern Europe and northern Africa, and the fact that Czech localities are rather isolated occurrences north of the Alps, we follow the treatment in Euro+Med Plantbase (Greuter 2006–2009), which considers the species as alien to the Czech Republic and assigns Czech populations to A. cretica subsp. cretica.

×Anthematricaria dominii (= Anthemis cotula × Matricaria chamomilla). A single plant was found at the Vltava river bank in Praha-Zlíčov in 1929 (Rohlena, PRC; Dvořáková in Slavík & Štěpánková 2004).

Artemisia alpina. One population was observed in Ujezd near Brno outside a garden in a partly mown lawn. Two young plants were found growing at a railway bank 80–100 m from the source population, suggesting that the species reproduced by seed at the locality, and died later due to summer drought (Čáp in Hadinec & Lustýk 2011).

Asparagus officinalis subsp. officinalis. This old cultural vegetable and medicinal plant has been widely cultivated in central Europe since the 16th century, and at the territory of the Czech Republic since the 18th century. It is naturalized in warm parts of the country. Some localities are remnants of cultivation in gardens or fields (Bělohlávková & Slavíková in Štěpánková 2010).

Avena sterilis. Two subspecies of the species given in Pyšek et al. (2002) are newly recognized in the country. Avena sterilis subsp. sterilis was planted in botanical gardens and nurseries in the 19th and the first half of the 20th century, with the first record of planting in the Kačina castle in 1836, from where it occasionally and temporarily escaped. The oldest records are from ruderal sites in Praha-Zlíčov (1922) and from a railway station Praha-Michle (1923). However, these records are not supported by herbarium specimens. The second subspecies, A. sterilis subsp. ludoviciana, was also formerly planted in botanical gardens, and occasionally found in waste places in Semily (1966 V. Jehlík, PRA), Prague (1968 Z. Kropáč, PRA) and Malý Budíkov near Humpolec (1965 A. Čábera, CB). Čábera (1967) published his find under the name A. strigosa (J. Zázvorka in Štěpánková in prep.).

Avena ×vivis (= A. fatua × A. sativa). Individual plants of this hybrid are occasionally found in the fields of A. sativa within the distribution range of A. fatua (J. Zázvorka in Štěpánková in prep.).

Berberis julianae. Self-sown young shrubs originated from a source population nearby were observed in a park plantation in Praha-Kláновice (50°05′42.2″N, 14°40′10.2″E) in 2010 (J. Sádlo).

Berberis thunbergii. A young shrub originated most probably from seed was found nearby a planting site in Stará Červená Voda, northern Moravia (50°19′44.9″N, 17°12′05.2″E) in 2011 (J. Sádlo).

Beta vulgaris Altissima Group. The annual weedy types that started to spread in the 1980s have been introduced with beet seed from southwestern Europe (Skalický & Pulkrábek 2006), where they originated through the pollination of cultivated sugar beet (Beta vulgaris Altissima Group) with the pollen of the wild B. vulgaris subsp. maritima or of weedy annual plants derived from some cultivars of the Altissima Group. For this reason, the assignment to the Altissima Group is a pragmatic solution, not fully reflecting the genetic nature of the plants concerned. A survey from 2006 revealed that “weed beet” occurred on 70% of farms over the Czech Republic growing sugar beet and on 4% of those its density exceeded 1000 plants/ha (Landová et al. 2010). The issue requires further study; the populations of weedy plants are now classified as invasive neophyte.

Bidens ferulifolius. Planted in flowerpots in towns and escaping from cultivation, growing in paving interstices and surviving temporarily, but not over winter (Mladá Boleslav and Náchod, P. Petřík; Bechyně and Prague, J. Sádlo). A vigorous population that was later destroyed by remodelling of the pavement was observed at the railway station in Jablonec nad Nisou, northern Bohemia, in 2006 (P. Petřík, pers. comm.).

Buddleja alternifolia. Several young shrubs up to 1.5 m tall, growing from seed, were recorded in ruderalized shrubland at abandoned factory yard in the Mostecká street, Chomutov, northern Bohemia (50°27′51.8″N, 13°25′12.7″E) in 2008 (K. Štajerová). The plants were present at this locality still in 2011 (J. Sádlo).

Buglossoides incrassata subsp. incrassata. A population of about 15 plants was observed at a railway station in Strážnice (distr. Hodonín, southern Moravia) in 2005 and first reported as Lithospermum arvense subsp. sibthorpiatum (Jongepier et al. in Hadinec & Lustýk 2006), but the revision of herbarium specimens (BRNU) revealed that the identification was erroneous (rev. J. Danihelka, conf. E. Zippel, Berlin). The population still occurred in the locality in spring 2012 when about 11 m long strip with 10–15% cover of flowering plants was recorded between the rails (J. Jongepier, pers. comm.).

Buglossoides incrassata subsp. splitgerberi. This subspecies was first reported from the Czech Republic as B. arvensis subsp. sibthorpiana by Clermont et al. (2003), based on the specimens issued as no. 1654 of Fl. Exs. Reipubl. Social. Čechoslov. However, Jongepier et al. (in Hadinec & Lustýk 2006) considered this record erroneous and assigned duplicates of that gathering to B. arvensis. The presence of this subspecies is now confirmed by
numerous herbarium specimens (rev. J. Danihelka, conf. E. Zippel) from both Bohemia and Moravia, collected mostly from ruderal sites and dry grasslands.

*Bupleurum pachyspernum.* The only find from the Czech Republic (1885 A. Oborny, PR) originates from the Dyje river valley near Znojmo, southern Moravia, and was reported by Snogerup & Snogerup (2001). It is considered here as a neophyte in accordance with the treatment for Austria (Fischer 2008) and in the Euro+Med Plantbase (Hand 2011).

*Buxus sempervirens.* Ongoing regeneration from seed is observed in the surroundings of planted shrubs in the Průhonice Park near Prague (J. Burda, pers. comm.). Several young shrubs were found in a natural ravine forest at Medník hill south of Prague, probably from self-seeding of shrubs planted near a cottage (2010 J. Sádlo).

*Calyptaris hederacea.* The species has been observed since ca 25 years ago growing on settling fields of a sugar refinery in Kojetín, central Moravia (Trávníček & Dančák 2011).

*Campanula lactiflora* is documented from one locality at Kladská (distr. Cheb, western Bohemia), where it occurs at the margin of a peat meadow (first collected in 1973, F. Grúll, BRNU as *C. latifolia*), probably as a consequence of plantings in the area of a hunting lodge built in 1877–1878 (Řehořek in Hadinec & Luxty 2009). Previously reported naturalized occurrence of this species in the former Czechoslovakia by Fedorov (1976) is doubtful and it is unclear on what data it was based (Řehořek in Hadinec & Luxty 2009).

*Caragana arborescens.* Reported as escaping from windbreaks in southern Moravia, where it is extensively planted (Tichá 2004).

*Capsella rubella.* A population consisting of tens of plants was found in 2006 in a camping site Babí hora near Hluk, distr. Uherské Hradiště, SE Moravia, probably introduced by foreign tourists. The species is native to southern Europe and reported from several countries north of its native distribution, e.g. Austria, Switzerland, Germany, Belgium and the UK (Jongepier in Hadinec & Luxty 2007).

*Caryx grayi.* A species occasionally planted in botanical and private gardens; three plants were found on a ruderal site at the railway station Zastávka u Brna, southern Moravia, in 2010. The plants did not persist until next year due to construction works at this site (Hrbáč in Hadinec & Luxty 2012).

*Centaurea carnatica.* A herbarium specimen collected in Hradec Králové was found in PRC (1914 K. Prokeš, Koutecký 2008).

*Centaurea xjavorkeae (= *C. nigrescens* × *C. oxylepis*). A hybrid involving the casual neophyte *C. nigrescens* was collected in 1933 near Litovel, distr. Olomouc (Novák, PRC; Koutecký & Štěpáněk in Slavík & Štěpánková 2004).

*Centaurea xextranea (= *C. jacea* × *C. nigrescens*). Another hybrid involving *C. nigrescens*, listed under the name *C. xhaizii* in the Flora of the Czech Republic, is documented with certainty from two localities but its occurrence is probable in other localities where mixed populations of both parents occur (Koutecký & Štěpáněk in Slavík & Štěpánková 2004).

*Centaurea transalpina.* Collected in Olřik nad Vltavou, southern Bohemia, around 1900 (K. Domin, PRC; Koutecký 2008).

*Cichorium endivia.* Escape from cultivation of about 40 plants in Brno-Lesná close to a bus stop was recorded in 1968 (Dvořáková in Slavík & Štěpánková 2004). In 2009, several tens of plants were recorded in an old field in the military training area of Bobolice, distr. Český Krumlov, southern Bohemia (Grulich in Hadinec & Luxty 2011). The species was most likely introduced to the country as a vegetable in the 16th century (Petráčková et al. 1982).

*Cirsium ×moravicium (= *C. arvense* × *C. rivulare*). This hybrid between an archaeophyte and a native species is known from one locality between the villages Ústí and Skalička, distr. Přerov, central Moravia (Bureš in Slavík & Štěpánková 2004).

*Convallaria majalis var. transcaucasica.* Planted in a hospital in Klatovy, western Bohemia, from where it spread into a nearby park and formed a viable population, which is still present. The introduction was by a local botanist M. Král in the 1970s (Čížek & Král 2009, Slavík & Zázvorka in Štěpánková 2010).

*Coreopsis lanceolata* was found in 1962 at Kunetická hora hill near Pardubice, eastern Bohemia, where it was surviving for several years, with most plants remaining sterile (Bělohlávková in Slavík & Štěpánková 2004).

*Corispermum declinatum* was collected in Praha-Stodůlky in 1960 (S. Hejný, PRC, det. J. Danihelka). The specimens come from the same locality as that of *C. pallasii*, treated in the Flora of the Czech Republic under the name *C. leptospermum* (Tomsíček in Hejny & Slavík 1990). This source notes that the collection included another species that remained unidentified, probably *C. squarrosum* (= *Agriophyllum squarrosum*) or *C. orientale*.

*Cotoneaster dielsianus.* Ongoing regeneration from seed is observed in the Průhonice Park near Prague (J. Burda, pers. comm.).

*Cotoneaster divaricatus.* A frequently planted species of the genus escaping from cultivation by seed dispersed by birds. It was recorded, for example, in Mikulov and Brno (J. Danihelka) or Praha-Klánovice (a fruiting

*Cotoneaster zabelii*. First reported from Černvír, distr. Žďár nad Sázavou, where several older shrubs and saplings grow on a rock above the Svatka river ca 100–150 m from the maternal shrub planted at a nearby house (Cáp in Hadinec & Lustyk 2007).

*Crocus tommasinianus*. The species was deliberately planted in the wild at Velká hora hill near Srbsko, Bohemian Mts, central Bohemia, before WWI, where it survived for several years, last observed in 1931 (Chrtek in Hadinec & Lustyk 2011 for details and references therein).

*Crocus ×crocosmiiflora*. Frequently planted hybrid, originated in cultivation, sometimes planted also in the wild or rarely escaping from cultivation (Chrtek in Štěpánková 2010).

*Cyperus glomeratus*. Rarely found escaped from cultivation (Kubát et al. 2002), first recorded in the Brdy Mts, central Bohemia, in 1895, later collected near Protivín, southern Bohemia, in 1947 and in Brno in 1965 (K. Kubát in Štěpánková in prep.).

*Darmera peltata*. The species was first found growing along a wet road ditch near Lukavice, western Bohemia, in 1960. The locality was later destroyed and the species was found again at the periphery of Klatovy town, in 1995 and later collected near Protivín, southern Bohemia, in 1947 and in Brno in 1965 (K. Kubát in Štěpánková in prep.).

*Digitaria ciliaris*. The occurrence of this species in the Czech Republic was first reported by Wilhelm (2009) who refers to a herbarium specimen from Podhůří, distr. Trutnov, collected on a decayed waste from cotton processing in 1908 (V. Cypers, BC). Another specimen from the same locality, collected one day later, is deposited at BRNU (no 5272, leg. V. Cypers). It needs to be noted that the name *Panicum ciliare* Retz., a basionym of the name *D. ciliaris*, repeatedly appears in herbaria and floristic literature from the Czech Republic since the first half of the 19th century, but based on morphological descriptions and numerous gatherings, the plants actually represent *D. sanguinalis* var. *pectiniformis*, which we consider a naturalized archaeophyte. The reference in Wilhelm (2009) to plants collected by V. Cypers is therefore the first record of this casual neophyte in the country (see Danihelka in Hadinec & Lustyk 2011 for details and references therein).

*Digitaria graveolens*. Thirteen localities from 2008–2009 are listed from the Czech Republic in a recent paper reporting it as a new species of the Czech flora (Raabe in Hadinec & Lustyk 2009). This Mediterranean species has been spreading rapidly in Central Europe, following the first reports at the beginning of the 1980s and 2000s in Germany and Austria, respectively, where it forms extensive stands in highway medians. In the Czech Republic, it was very abundant along the D1 highway Prague – Brno already in 2008, forming large stands close to Brno and Velké Meziříčí (Raabe in Hadinec & Lustyk 2009). In 2011, it was seen at other 10 sites between km 27 and km 106 (U. Raabe, pers. comm.). At present it spreads further southeastwards to Bratislava, and it is also recorded from the D11 highway (J. Rydlo, pers. comm.) and the Nymburk district, central Bohemia (F. Krahulec, pers. observ.). The species is thus classified as naturalized even though the first documented record from the Czech Republic is very recent.

*Egeria densa*, planted in aquaria, was observed two times in the wild: in a pond in the Kinského sady park in Prague in 1991, and in a village pond in Borek near České Budějovice, southern Bohemia. The finds are most likely due to deliberate release; the plants do not survive winter in local conditions (Kaplan in Štěpánková 2010).

*Elaeagnus commutata* was planted on a spoil heap Antonín in the Sokolov coal mining area, northwestern Bohemia, during rehabilitation activities in the first half of the 1970s (Dimitrovský 2001) and spread over an area of several hectares, first along roads but gradually also elsewhere, forming dense stands in places (P. Krása & V. Grulich, pers. comm.).

*Eragrostis pectinacea*. The species was first collected in a botanical garden in Olomouc (1937 O. Leneček, PRC) and one tussock was observed in Pardubice, eastern Bohemia, in 2000–2001 (P. Špryňar, PRC). It was reported as a new alien species in the Czech flora based on a thorough revision of herbarium collections of the genus (Špryňar & Kubát 2004).

*Eragrostis pilosa* was traditionally considered as native based on a near-natural character of the locality from which it has been known since the beginning of the 20th century, but this view was recently reconsidered based on a revision of the genus in the country (Špryňar & Kubát 2004). The species was first collected at Znojmo-Hradiště, southern Moravia, in 1902 (A. Wildt, BRNM) where it still grows, and reported from several other localities in warm regions, including slaughter house in Praha-Holešovice where it was surviving for 30 years. It is still included among Red List species as a critically endangered (Holub & Procházka 2000), based on the Hradiště locality. We follow the opinion of Špryňar & Kubát (2004) and consider it as a naturalized neophyte.
**Euphorbia agraria**. A single plant was found growing on abandoned valley terraces close to Komoroňany, distr. Vyškov, in 2005, and disappeared by 2008 when the grassland was mown. The find represents the first report not only for the Czech Republic but the whole of Central Europe (Čap 2008, Čap in Hadinec & Lustyk 2009).

**Euphorbia myrsinites** was found in 1998 growing on garden waste at an abandoned quarry on Svatý kopček hill near Mikulov, southern Moravia (J. Danihelka). In 2009, several tens of plants were found in a sand pit in Tasovice near Znojmo, southern Moravia. Most likely, deliberate planting of the species in the wild was followed by its proliferation by seed (J. Sádlo).

**Euphrasia salisburgensis** and *Gentianella obtusifolia* subsp. norica. These two species were most likely deliberately introduced to the Rýchorý range, Krkonoše Mts, at the end of the 19th century (Štursa et al. 2009), and repeatedly collected during the first half of the 20th century mostly around the Rýchorská studánka spring. The idea of deliberate introduction into the wild is supported by the species being not reported as a part of local flora by botanists working in the area in the 19th century (A. F. Pax, R. Traxler).

**Fallopia × convolvuloides**. A hybrid between an archaeophyte *F. convolvulus* and a native species *F. dumetorum*, occasionally found where both species grow together (Chrt in Hejny & Slavík 1990).

**Ferulago confusa** was collected at two localities: in an oak forest in the Koda Nature Reserve near Tetín, distr. Beroun, central Bohemia, in 1998 (one plant), and in a dry grassland in the Kamenný vrch Nature Reserve in Brno-Starý Liskovec in 2002. The species still occurs at the latter site, with 2–3 flowering plants observed every year (O. Rotreklová, pers. comm.). It is not cultivated in the Czech Republic, except perhaps in some botanical gardens, and it is not reported as escaped from cultivation in the neighbouring countries. The way of introduction is therefore unclear, and given that both localities were discovered at about the same time, deliberate sowing cannot be excluded (Rotreklová & Rehořek in Hadinec & Lustyk 2009).

**Filago pyramidata** was collected at two localities, Olomouc and Olomouc-Černovír in 1833 and 1860, respectively (both specimens at W), and not observed since then (Wagenitz 1965). This corresponds to the fact that the species’ native distribution was more extensive until the 19th century, allowing for introductions to the Czech Republic, but it has been retreating since then (M. Štech in Slavík & Štěpánková 2004).

**Gaillardia × grandiflora** is a commonly cultivated ornamental hybrid, occasionally found escaping (Bělohlávková in Slavík & Štěpánková 2004). It was recorded in Mikulov, southern Moravia, where the plants seeded for two years in 2003–2004 at the foot of a wall, and in Břeclav-Poštorná, southern Moravia (2003 J. Danihelka, MM).

**Galium murale**. Five fruiting plants were recorded at the Albertov university canteen entrance in Prague in 2009, eight plants in 2010 and two plants in 2011. In Europe the species was up to now only reported as an alien from the UK and Belgium (Prančl in Hadinec & Lustyk 2012).

**Geranium purpureum** was first recorded at a railway station Hrušovany u Brna in 2005. Three years later it was found at all stations between Hrušovany and Brno (Růžička & Koblížek 2009). Further spread is likely in the near future.

**Gleditsia triacanthos**. The species occasionally occurs in near-natural vegetation. While it is not clear whether older trees are cultivation remnants or established spontaneously, a massive occurrence of seedlings was documented from an exposed bottom of the Prostřední rybník fishpond near Lednice, southern Moravia (2008 J. Danihelka, BRNU).

**Gratiola neglecta** was recorded at two sites near Lázně Bohdaneč, eastern Bohemia, in 2002, and at one site in the surroundings of Blatná, southern Bohemia, in 2008 (Šumberová & Ducháček 2009).

**Helianthemum nummularium** subsp. nummularium. Occasionally planted as a garden ornamental, reported to escape in Olomoučany near Brno (Hrouda in Hejny & Slavík 1990).

**Herniaria incana**. The species was recorded in a mown dry grassland near Hřímovice, central Bohemia, in 1986 (Hlaváček 1989). It reproduced by seed and persisted in the locality until the beginning of the 1990s (Hlaváček & Pyšek 1992). Later it started to retreat due to changes in the management of the site (R. Hlaváček, pers. comm.).

**Hieracium mixtum**. A population of a flowering maternal plant and several juveniles was found growing on a stony slope along a hiking trail to Mt Praděd, Hrubý Jeseník Mts, at 1355 m a.s.l., in 2006. By 2010, the population increased and one of the juveniles was also flowering. The species is a triploid apomict not requiring pollination
for seed production, and is rarely planted as an alpine plant in rockeries, and traded by garden centres. Its occurrence most likely results from deliberate planting or sowing in the wild (Kocián & Chrtek in Hadinec & Lustyk 2011).

Hordeum brevisubulatum. A herbarium specimen collected in 1974 (M. Dvořáková?, BRNU 605154) on a waste place in a textile factory Brunka in Humpolec, distr. Pelmřimov, was identified as the first record of this species in the country. It was almost certainly introduced with wool of Soviet origin, most likely from southern Siberia or central Asia (Danihelka in Hadinec & Lustyk 2009).

Hyacinthoides hispanica was recorded escaped from cultivation near a fishpond in Prague and in a forest near Mašov, eastern Bohemia, in 2007 and 2008, respectively (Trávníček 2010, Hadinec & Lustyk 2012). Another plant was found in the Heršťn Nature Reserve near Kdyně, western Bohemia, in 2009 (P. Petřík, pers. comm).

Hypericum annulatum. A population of about 20 plants was recorded on a power plant fly ash heap near railway station at Oslnavy, distr. Brno, in 2008 (Sutorý 2010a, b, Hadinec & Lustyk 2012).

Koelreuteria paniculata. Copious regeneration from seed was observed in park plantations in Brno and Lednice, both in 2009 (J. Sádlo).

Lamium × holsaticum. A hybrid of the archaeophyte L. album with the native L. maculatum, assumed to occur rather frequently near the populations of its native parent (Dvořáková in Slavík 2000).

Lathyrus hirsutus. The species was considered native in the Flora of the Czech Republic (Chrtková & Bělohlávková in Slavík 1995), but its native distribution range in southern Europe, character of habitats and namely the absence from old floras (e.g. Čelakovský 1868–1883) are arguments against its native status; we suggest it be classified as a neophyte but its status requires further study (see also Hadinec & Lustyk 2011).

Legousia pentagonia. Recorded in the Czech Republic for the first time in 2005, when several flowering plants were found on a forest clearing along a road near Lanžhot, southern Moravia, together with two plants of another casual neophyte, Gilia achilleifolia. The species was probably introduced to the site with forestry vehicles (Řehořek & Lososová in Hadinec & Lustyk 2009).

Lilium bulbiferum. Although some authors consider its localities in southern Bohemia as a margin of its native distribution, we classify the species as an archaeophyte, following the recent treatment in Flora of the Czech Republic (Hrouda in Štěpánková 2010).

Lilium candidum. A frequently planted species, occasionally surviving as a cultivation relic, or growing in places with deposited garden waste (Hrouda in Štěpánková 2010).

Lolium × hybridum. A hybrid between a neophyte L. multiflorum and the native L. perenne is reported to occur by Kubát et al. (2002).

Lonicera periclymenum is occasionally reported in the literature, but without the character of occurrence specified. Therefore it is in most cases difficult to decide whether the reports relate to surviving, originally planted shrubs, cultivation relics or escape. Extensive clonally spreading stands were recorded in ruins of the Ronov castle near Česká Lípa (50°37'13.3"N, 14°24'52.1"E) in 1994 (J. Sádlo) and in a wet forest margin near Doksy (50°39'12.2"E) in 2010 (J. Sádlo), both northern Bohemia.

Lotus ornithopodioides is occasionally found on fodder plots in game preserves in southern Moravia, growing from seed most likely originating from Fodder Research Institute in Troubsko near Brno. It is documented by herbarium specimens from two sites: (i) Mikulov: Bulharská obora game reserve, fodder plot between a water hole and path along the fence, ca 4.1 km ENE–E of the town church (1997 J. Danihelka, MMI, det. T. Vymyslický); (ii) Lanžhot: a small forest meadow south of the bend of the Iklínská cesta forest road, 3.7 km SSE–S of the church (1996 J. Danihelka, MMI).

Malus baccata. Ongoing regeneration from seed is observed in the whole area of the Průhonice Park near Prague (J. Burda, pers. comm.).

Malus fusca. A young flowering shrub grown from seed was found in 2004 near Telnice, distr. Brno. The species was not observed in cultivation in the wider surroundings of the locality, suggesting probable dispersal by birds. Seeds collected at the locality germinated easily (Řehořek in Hadinec & Lustyk 2009).

Malus pumila. A species of unclear origin, introduced to Europe from the Southern Caucasus where it does not, however, occur in the wild (Dostálek in Hejný & Slavík 1992). It is known from the territory of the present Czech Republic since 1852. At present it is mostly planted as a rootstock for M. domestica and occasionally reported as escaped in the floristic literature from various parts of the country (Křivánek 2008).

Malva sylvestris var. mauritiana. The taxon was listed in Pyšek et al. (2002) at the species level, now both varieties occurring in the country are included (Appendix 2). This old variety of unclear origin is occasionally planted as a medicinal or ornamental herb and temporarily escapes from cultivation (Slavík in Hejný & Slavík 1992).
Matricaria chamomilla. The species is considered as an archaeophyte in Central Europe, following the treatment in the Flora of the Czech Republic. It is planted as a medicinal herb and is common throughout the country as a weed on arable land or at ruderal sites (Kubát in Slavík & Štěpánková 2004).

Matricaria chamomilla × Tripleurospermum inodorum. An intergeneric hybrid between two archaeophytes, so far only reported from Germany and the Czech Republic. Several plants were collected at two localities in Prague in 1929 (Rohlena 1930, Kubát in Slavík & Štěpánková 2004).

Meconopsis cambrica. A frequently planted ornamental species, reported to escape easily from cultivation. It was recorded spreading in an abandoned garden in Zahrady, distr. Děčín, northern Bohemia, in 2000 (Kubát in Hřítel et al. 2002, Hadinec et al. 2003).

Miscanthus sacchariflorus. Several tussocks grown from seed were observed in a garden allotment in Ostrá, distr. Nymburk, central Bohemia, in 2003, and one young plant on a garden waste in a quarry near Velká Vápenná, Jeseníky Mts in 2010 (J. Sádlo).

Morus alba is reported in literature (Křivánek 2008), but in most cases it is difficult to decide whether the reports relate to surviving, originally planted trees, cultivation relics or escapes. Regeneration by seed is, however, reported from Slovakia, where the species is classified as naturalized (Medvecká et al. 2012).

Muscardi armeniacum was reported as likely to escape from cultivation but not observed as such in the Flora of the Czech Republic (Hrouda in Štěpánková 2010). However, it was recorded in many localities in Prague and Mělník, central Bohemia (both J. Sádlo), and Brno and Mikulov, southern Moravia (both J. Danihelka).

Muscardi botryoides is an archaeophyte with scattered distribution in the past, but not observed since the last record in 1995 (Hrouda in Štěpánková 2010).

Opuntia polyacantha is a taxonomically complex species, also reported under other names in the horticultural literature, e.g. O. erinacea var. utahensis (Biba 2007). Here these two names are synonymized following a flora from the species’ native range (Pinkava 2003). It is recorded from several localities in warmer regions, e.g. Lovoš hill near Litoměřice, northern Bohemia, Prague and Brno and their surroundings. Populations range from those of seedlings (Průhonice Park near Prague) to those of polycormons with estimated age of 20 years in the Skalky u přehrady Nature Reserve near Brno-Bystrc (L. Tichý, pers. observ.). It is assumed to have been deliberately planted in these localities (Hadiniec & Kubát in Hadinec et al. 2004). The first observation of such plants comes from the Dalejský profil Nature Reserve in Prague in 1997 (Šprýňar et al. 1998). The other species of the genus, O. phaeacantha, listed in Pyšek et al. (2002) is reported from Slanská hora hill in the town of Slaný, central Bohemia, surroundings of Prague, the České středohoří hills, northern Bohemia, and the Pavlovské vrchy hills, southern Moravia, assumed to persist following deliberate planting (Kubát et al. 2002, Pyšek et al. 2002).

Paeonia lactiflora escapes from cultivation in gardens, persists in abandoned nurseries and garden allotments, and on rubbish tips from garden waste; it regenerates vegetatively from rhizome segments. It was recorded as escaped e.g. in Praha-Kbely in 2011 (J. Sádlo).

Pennisetum alopecuroides. One flowering plant was found in Praha-Satalice, on stairs of a house, in 2002 (J. Sádlo).

Physalis pubescens. A single plant was recorded on a soil heap in Zlatá Koruna, distr. Český Krumlov, southern Bohemia, in 2001, but no longer found when the locality was revisited in 2002 (Lepší 2005).

Pimpinella peregrina. A population of this species scattered along about 300 m long strip on a ruderal site was recorded in Ústí nad Labem in 2011 (Nepraš et al. 2011). Its introduction was probably linked with recent remodelling of a railway corridor. Its spread in the neighbouring Saxony, observed since the 1990s, is attributed to grass seed used for revegetation following building activities (Nepraš in Hadinec & Lustyk 2012). Further spread in the Czech Republic thus cannot be excluded.

Pinguicula crystallina subsp. hirtiflora and P. grandiflora subsp. rosea. Both taxa were recorded on a tuft cascade in a forest near Tichá in the Beskydy Mts, distr. Nový Jičín, northern Moravia, in 2006, where it was probably deliberately planted. The population of the former taxon has considerably spread in the locality since then (A. Veleba, M. Chytrý).

Podophyllum hexandrum. Found at a forest margin by the Nový Herštejn castle near Kdyně (49°24'45.8"N, 13°04'00.2"E), western Bohemia, in 2009 by P. Slovák, close to an abandoned garden (P. Petřík, pers. comm.).

Pontederia cordata is occasionally planted as an aquatic ornamental in garden ponds, and recorded from several localities outside cultivation in 2004–2007. These occurrences are mostly due to deliberate planting in the wild, with plants surviving as cultivation relics. Plants found in the Labe river near Chvalovice (river km 62.91), central Bohemia, were obviously dispersed to the site by water and observed in two subsequent years, 2006 and 2007 (Kaplan in Hadinec & Lustyk 2009).

Potentilla adscharica. The species was first collected escaped in the Botanical garden of Charles University in Prague in 1947, then repeatedly at the then unfinished Prague – Brno highway in 1950–1956; both finds were
probably related to plants spreading from the botanical garden, and they are only two records of this species in Europe (Soják 2007).

**Potentilla radiata.** Repeatedly collected in the Průhonice Park near Prague in 1920–1926; the occurrence has not been confirmed since then (Soják 2007).

**Primula rosea.** A Himalayan species recorded for the first time in the Czech Republic at two sites in the Praděd Nature Reserve, Hrubý Jeseník Mts, in 2005. It is likely that this popular garden ornamental was deliberately planted in a spring fen where it occurs (Kočí in Hadinec & Lustyk 2007).

**Ptelea trifoliata** is reported growing along roads in Bruntál, northern Moravia (Opravil 1961). A fruiting shrub was also seen at the fence of the Michelská plynárna gasworks in Prague in 1964 (Škalická & Svoboda 1971).

**Pteris multifida.** One plant was found growing in a wall crevice in Prague in 1998, but it was destroyed next year during the facade renewal (Ekrt in Hadinec & Lusty 2011).

**Pulmonaria rubra.** Several populations were found scattered in different habitat types along ca 2 km of the Všenorský potok stream near Všenory, distr. Praha, in a woodland valley. The species was first collected in 2001, then again in 2002 but at the time of the first collection it was already growing at that site for some time (V. Větvička, pers. comm.). The species is very rarely planted in the Czech Republic as a garden ornamental. It is very likely that the escape from cultivation is related to a former experimental gardening centre, used as an acclimation garden of the Institute of Botany AS CR, which is located up the stream (Hadinec & Rydlo in Hadinec et al. 2004).

**Pyracantha coccinea.** A popular ornamental shrub, recently found escaping from cultivation with increasing frequency in urban shrubland and grassland, ruderal sites, usually spread by seed to a short distance (up to 100 m) from cultivated plants. Numerous localities were recorded in Prague in 2002–2012 (J. Sádlo). One shrub 4–5 m tall was also found in a shaded forest near the Koněpruské jeskyně caves, distr. Beroun, central Bohemia, in 2008 (R. Hlaváček, pers. comm.).

**Rhaphonticum carthamoides.** The species started to be planted as a medicinal plant in the 1980s. It was first recorded escaped from cultivation in 1991 in a road ditch near Vlkava, central Bohemia, not far from a field where it was planted. The second record, from 2003, refers to individual plants surviving on an abandoned field in Velký Osek, central Bohemia, after the cultivation has ceased (Rehořek in Hadinec et al. 2004).

**Rheum officinale.** Five localities are reported in the Novohradské hory Mts, southern Bohemia, one of them has been surviving since the 1980s (Lepší et al. 2006). Other localities found recently in mountainous areas of northern Bohemia (Šída in Hadinec & Lustyk 2008) make further spread of the species likely.


**Rhodotypos scandens.** A locally naturalized population has been observed since the early 1990s in the Boří les forest introduced to cultivation in the region probably in the 1920s, when former pastures south of the Prostřední rybník fishpond were afforested mainly with introduced species (J. Uhr, pers. comm.). This is the first case when the species became locally established in Europe; so far it is only reported as casual from Belgium and Hungary (DAISIE 2009), as well as from Vienna in Austria (Fischer 2008).

**Ribes sanguineum.** Several tens of flowering shrubs were found along a tourist path in a spruce and larch plantation close to a chalet settlement near Dolany, distr. Olomouc, central Moravia. As the species increasingly appears on sale in garden centres, its spread by birds is likely. Plants growing in the locality probably belong to some of the numerous garden cultivars (Hadinec & Prach in Hadinec & Lustyk 2008).

**Rodgersia podophylla** survives as cultivation relic in parks for many decades, e.g. in Průhonice or Vrchotovy Janovice. Seeds do not germinate and plants spread only vegetatively (Sekerká 2009).

**Rosa multiflora** is reported as escaping from windbreaks in southern Moravia (Tichá 2004). Shrubs most likely established from seed were repeatedly observed at urban sites in Prague (J. Sádlo, pers. obs. 2009 and 2012).

**Rudbeckia fulgida** is a garden ornamental once documented as temporarily escaped from cultivation on a ruderal site at the Pustý rybník fishpond near Blatná, southern Bohemia (Deyl & Skočdopolová-Deylová 1989).

**Rumex patientia** × *R. tianschanicus* is a hybrid originated in cultivation in the Ukraine and planted as a biofuel crop in the Czech Republic since the 2000s, usually referred to as *Rumex ‘Uteuša’*. Probably the first record outside cultivation was a single sterile plant at the western shore of the Rozkoš water reservoir, eastern Bohemia, in 2005, ca 3 km from the nearest planting plot (F. Krahulec). Since then it has been repeatedly reported as escaping from cultivation in other places elsewhere.

**Rumex longifolius** subsp. *sourekii*. Pyšek et al. (2002) listed only the species *R. longifolius* without indication of subspecies. Now both subspecies occurring in the country (subsp. *longifolius* and subsp. *sourekii*) are included.
This subspecies occurs in disturbed habitats at higher altitudes. In the 1990s it was locally common (Krkonoše Mts, Jizerské hory Mts) and spreading (K. Kubát in Hejní & Slavík 1990).

*Salix melanopsis* is locally naturalized at the Nové Mlýny water reservoirs, southern Moravia, where it was planted to prevent bank erosion in 1984–1992. It spreads by vigorous root suckers, but rooting of branches dispersed by water was also observed. The species was recorded on 10 out of 16 islands investigated and on the upper dam of the middle reservoir of Nové Mlýny. Plants cultivated in the Czech Republic are a single clone (Úradníček 2004).

*Salix cordata*. Originally reported as a find of a rooted branch at the Rovenský rybník fishpond under the name *Salix ‘Americana’* (Krahulec 1975). A clone of this species (det. J. Koblížek) persists at one site in Česká Skalice, eastern Bohemia, since the 1960s, most probably as a cultivation relict; the population is maintained by rooting.

*Santolina chamaecyparissus* is occasionally planted in gardens and reported to escape rarely and temporarily, e.g. near Ledeč nad Sázavou, eastern Bohemia (Bělohlávková in Slavík & Štěpánková 2004).

*Sarracenia purpurea*. About 10 plants were recorded at the Ránské fishpond near Křižánky in the Žďárské vrchy Mts, eastern Bohemia, in 2011, having survived winter from the previous year. The plants were assumed to have been deliberately planted in the wild and since the locality is in a protected area, nature conservation authorities planned their eradication when the species was found. As the information appeared on the internet (http://www.novinky.cz/domaci/229243-na-vysocine-se-objevila-americka-masozrava-rostlina.html), the identification based on a photograph was possible. The species was also observed to survive winter and produce seedlings in a peaty site in a private garden in Liberec (L. Sekerka, pers. comm.). In the Borkovická blata peatbog near Soběslav, southern Bohemia, a single plant was planted in the wild, survived winter for several years and produced numerous seedlings before it was eradicated (M. Štech, pers. comm.).

*Sasa palmata* ‘Nebulosa’. Two dense stands, the larger one of about 150 m², were found in Praha-Hostišov (50°07.25.1"N, 14°24.06.4"E) in 2012, probably resulting from former cultivation and subsequent vigorous clonal spread (J. Sádlo).

*Scilla forbesii*. The species is often planted and known to escape from cultivation in some botanical gardens and parks, first reported in the Podzámecká zahrada garden in Kroměříž in 1934 (H. Zavřel, BRNM, PR). Two confirmed records in the wild come from the surroundings of Prague, near Lhota in 1998 and in the Miličovský les wood in 2000. It is likely that several herbarium specimens from the second half of the 20th century, determination of which was not possible due to collections late in the season, also relate to the species (Trávníček 2010, Trávníček in Štěpánková 2010).

*Scilla sardensis* is occasionally planted and recorded as escaped in two localities. A population of ca 100 plants was first observed in the castle park in Otín, western Bohemia, in 1965; by 2004 it has increased to 500–600 plants spontaneously occurring in the park (Král et al. 2004a). It was reportedly planted in the wild in the Průhonice Park near Prague (Blažek 1972), and recorded spontaneously growing in several other localities such as Luděřov, central Moravia, and the university botanical garden in Olomouc (Trávníček 2010, Trávníček in Štěpánková 2010).

*Senecio xhelwingii*. A hybrid between the neophyte *S. vernalis* and the archaeophyte *S. vulgaris* is rarely found in populations of parental species (Gruulich in Slavík & Štěpánková 2004).

*Scolymus maculatus*. Collected in 1969 at a rubbish tip in the former loam pit (“Kohnova cihelna”) below Červený kopec hill in Brno, southern Moravia (F. Grüll, BRNU, det. J. Danihelka). It was erroneously determined as *Carthamus lanatus* and published under this name by Grüll (1979).

*Sorbus austriaca*. A hybrid between the neophyte *S. vernalis* and the archaeophyte *S. vulgaris* is rarely planted in gardens and reported to escape rarely and temporarily, e.g. near Ledeč nad Sázavou, eastern Bohemia (Bělohlávková in Slavík & Štěpánková 2004).

*S. × helwingii* was formerly considered native but the plants actually represent another species. A taxonomic revision revealed that *S. austriaca*, with the native distribution range from the Pyrenees to the Alps, has been planted in the Czech Republic since at least 1966 as a garden ornamental and alley tree, and rarely escapes from cultivation. So far it has been documented from two localities in central Bohemia: Průhonice and Benešov. A population of tens of young individuals up to 2–3 m tall was found growing along a tourist path in a woodland from cultivation. So far it has been documented from two localities in central Bohemia: Průhonice and Benešov.

*S. iranica*. It is unlikely that the species was planted nearby as it is not used as a garden ornamental or medicinal plant in the Czech Republic (Řehořek et al. in Hadinec & Lustyk 2009).
Symphyotrichum laeve × S. lanceolatum is a stabilized hybrid similar to taxa known from the native range in North America. Plants were so far only collected in Moravia: around Brno, Vyškov, Frydek, and in the Moravian karst (Kovanda & Kubat in Slavik & Stošáková 2004).

Tagetes tenuifolia is reported as not known to escape from cultivation in the Flora of the Czech Republic (Bělohlávková in Slavik & Stošáková 2004), however, it was recorded in Nová Ves u Bakova in 2009 (J. Sádlo).

Thuja occidentalis. Young trees were recorded in Praha-Satalice, planted trees also occasionally regenerate on cemeteries and in villages (2012 J. Sádlo). It was also observed to regenerate, mostly from seed, near planted individuals in the Průhonice Park near Prague (J. Burda, pers. comm.).

Trachyspermum ammi was recently reported as a new alien species for the Czech Republic based on a herbarium specimen collected in Ústí nad Labem-Svádov on a sandy bank of the Labe river in 1903, and recently identified by M. Marek. The species probably originated in cultivation (Hadinec & Lustyk 2012).

Trifolium alpinum and T. badium. The species were most likely deliberately introduced to the Rychory Range, Krkonoše Ms, at the turn of the 19th century (Štursa et al. 2009). Trifolium alpinum was collected once in 1919 (Kubat in Slavik 1995). T. badium repeatedly during the first half of the 20th century mostly around Rychorská studánka spring, where it still survived at the end of the 2000s (F. Krahubec). The hypothesis of a deliberate introduction to the wild is supported by these species being not reported by the 19th century botanists working in the area (A. F. Pax, R. Traxler).

Trifolium vesiculosum was collected in 1989 in a field near Troubsko, distr. Brno, where it was previously planted as a genetic resource for fodder production (R. Řepka, BRNU), and escaped in Louky, distr. Zlín, northern Moravia, in 2009 (Řehořek in Hadinec & Lustyk 2012).

Viola septemloba. An abundant self-sustaining population of the species was found at the Central Cemetery in Brno-Bohunice, in the part with soldiers’ graves from the 1920s. It was first collected by K. Sutorý (BRNM) in 2003. This author suggests that since the species is not known as planted in Europe, it might have been introduced by legionnaires returning from Russia via North America to the former Czechoslovakia after WWI (Sutorý in Hadinec & Lustyk 2008).

Vaccinium corymbosum. Hundreds of plants originated from seed were recorded in a peaty forest in the Borkovická blata peat bog near Mažice, southern Bohemia, in an abandoned planting site (2011 J. Sádlo).

Viburnum rhytidophyllum. Several young shrubs were found growing in a hedgerow in Brno-Řečkovice in 2011, resulting from natural regeneration of two large shrubs grown nearby (J. Danihelka), and in an abandoned garden in Průhonice (J. Sádlo).

Changes of immigration status

Residence time: neophytes reclassified as archaeophytes

Most changes to the residence time status are based on the sources that the authors of the original catalogue were not aware of, or that appeared since the publication of the original catalogue (Pyšek et al. 2002). This concerns namely extensive archaeobotanical research focusing on thorough analysis of archaeological sites in several parts of the Czech Republic, carried out by V. Čulíková (Most, Prague, Česká Lípa, Libice nad Cidlinou, Čáslav, Opava; summarized in Čulíková 1986, 1994, and reported in numerous papers referred to below) and E. Opravil (e.g. Opravil 1980, 1986, 1993, 1994). This research provided evidence of the medieval presence of a number of
taxa previously considered as neophytes at the territory of the Czech Republic: these taxa need to be reclassified as archaeophytes.

*Allium cepa* was part of medieval diet and is sporadically documented by archaeobotanical finds so far. A find from 1438 at Kozí Hrádek (distr. Tábor) is documented, and sporadic onion seed from the High Medieval come from Opava and Jihlava (Čulíková 2000; see also Čížek 1994).

*Anthriscus cerefolium* var. *cerefolium* was cultivated as a vegetable since the Medieval (Slavík in Slavík 1997a) at it was escaping in the past (Koutecký in Hadinec et al. 2004).

*Arrhenatherum elatius* was already reclassified as an archaeophyte due to the lack of clear evidence for its introduction only in the Modern Period (Chytrý et al. 2005, Sádlo et al. 2007). This reclassification is supported by archaeobotanical evidence from the work of Čulíková (1999) who found five caryopses in Libice nad Cidlinou in the material from the mid 10th century. Other archaeobotanical finds of *A. elatius* come from the 16th century (Čulíková 1995b, 2002). Recently Poschlod et al. (2009) argued that the neophyte status is more appropriate for *A. elatius* var. *elatius* because the medieval archaeobotanical records refer to *A. elatius* var. *bulbosum*, native to southern and southwestern part of Central Europe (Conert 1998: 231–232). There are a few records of the latter from the Czech Republic (Dostál 1989: 1381). However, as M. Dvořáková (in Štěpánková in prep.), who treated the species for the Flora of the Czech Republic, could not find any herbarium specimens of var. *bulbosum* collected in the country, we include *A. elatius* only at the species level and consider it as an archaeophyte. The issue, however, requires further study.

*Atriplex hortensis* was used as a vegetable and medicinal plant in the Medieval, and its achenes were found from archaeobotanical sites in the town of Most, northern Bohemia, dated to the 13th and 14th centuries (Čulíková 1981, 1995b).

*Camelina microcarpa* was repeatedly documented by archaeobotanical studies to be regularly present at several archaeological sites (Prague, Most, Libice nad Cidlinou, Opava) since the 10th century (Čulíková 1998a, b, 1999, 2001a, b, 2002, 2005, 2006, 2009, 2010).

*Camelina sativa*. There are rare archaeobotanical finds of the seed of this species in medieval diet. Čulíková (2000) points out that while they do not provide unequivocal proof of its cultivation, the species is considered a traditional oil plant (Čulíková 2000).

*Chenopodium follicosum* was recorded in a fill of a waste pit from the 14th century in Most, northern Bohemia (Čulíková 1981).

*Citrullus lanatus*. Three localities (in Prague and Opava) are reported in the CZAD (Archaeological Institute ASCR 2011).

*Coriandrum sativum*. Three localities (in Prague and Opava) are reported in the CZAD (Archaeological Institute ASCR 2011).

*Cucumis melo* was reported to occur in sporadic finds from Bohemia and Moravia dating back to the late Middle Ages (Čulíková 2000).

*Cucumis sativus*. The earliest record of this species comes from the 9th–10th century Prague (Čulíková 2001a); it was further documented in a number of archaeobotanical studies in northern Bohemia and Prague (Čulíková 1981, 1995b, 1997a, 2000, 2001a, b, 2002, 2005, 2010).

*Daucus carota* subsp. *sativus*. Historical evidence suggests that it has been planted in Central Europe since the High Medieval; the region of carrot planting in Europe extended during that period from the southwest to the north and east, with reports from neighbouring Poland in the 14th century (Stolarczyk & Janick 2011).

*Dipsacus satisus* is reported from the Medieval (Opravil 2000). It is also recorded from that period in Germany (Knörzer 1984) and Great Britain (Ryder 1994).

*Elsholtzia ciliata* was recorded in a fill of a waste pit from the 14th century in Most, northern Bohemia (Čulíková 1981, 1995b).

*Gallega officinalis*. Pollen of this species was recorded in an archeobotanical profile from Libice nad Cidlinou, central Bohemia, from the Early Medieval (R. Kozáková, unpublished).

*Fagopyrum esculentum* was first documented from the turn of the 9th century in Prague (Čulíková 1998a, 2000) as well as from several later medieval sites (Čulíková 1987, 1995b, 2002).

*Ficus carica*. Its fruits are considered to be imported already in the Medieval, although it may have been occasionally cultivated in warmer regions of the country (Čulíková 2000; see also Čížek 1994). The oldest documented record comes from Prague already from the 9th century (Čulíková 1998b, 2001a). Achenes were usually found in abundance at each locality subject to archaeobotanical research (Čulíková 2000) throughout the Medieval (Čulíková 1981, 1987, 1995b, 1997a, b, 1998a, b, 2001a, 2002, 2003, 2005, 2009, 2010).
Glaucium flavum was recorded, as a seed, in samples from Prague dated to the 9th–10th centuries (Čulíková 2001a).

Iris × germanica and Iris × sambucina. Planted in central Europe since the Medieval (Jäger et al. 2008).

Lathyrus sativus. Palaeobotanical evidence suggests that the species was planted in the Bronze Age, recorded from Dobšice, southwestern Moravia (Kočár & Dreslerová 2010).

Lens culinaris was recorded from the medieval period in the Prague Castle (Čulíková 2001b). Two seeds were also recorded in a 13th century sample from a well in the Prague Castle. Although the species has been planted in the region since prehistoric times it is rarely recorded in medieval archaeobotanical samples (Čulíková 2012).

Levisticum officinale was present in several archaeobotanical samples from Most and Prague from 13th–15th centuries (Čulíková 1981, 1987, 1995b, 2002). It is supposed that it has been more widespread in the Medieval than indicated by the frequency of its finds (Čulíková 2000).

Myrrhis odorata was present in the Medieval from Prague (Opravil 1986) and was well known from Central Europe in that period (Harvey 1984).

Prunus cerasifera was reported by Čulíková (1995b) from Most, which is not unambiguous evidence as the dating of this site extends until the 16th century, but there are several records from the High Medieval in the CZAD (Archaeological Institute ASCR 2011).

Rapistrum rugosum. The species was recorded from the medieval period at Mikulčice, southern Moravia (P. Kočár, pers. comm.). There are two subspecies distinguished in the Czech Republic, subsp. rugosum and subsp. orientale, both up to now considered as neophytes first recorded in the Czech Republic in 1850 and 1940, respectively (Smejkal in Hejný & Slavík 1992). As they cannot be separated based on archaeobotanical evidence, we use this find as the reason for classifying subsp. rugosum as an archaeophyte.

Salvia officinalis was recorded in a fill of a waste pit from the 13th–14th centuries in Most, northern Bohemia (Čulíková 1981, 1995b; see also Čížek 1994).

Satureja hortensis was recently confirmed with certainty from a 13th century sample in Čáslav (Čulíková 2011b). Until now it was missing from the largest medieval sampling site in Most (Čulíková 1994) and previous records did not allow unambiguous identification, with the exception of a fill of a waste pit in Opava from the 15th century (Čulíková 2011a) and records from the early post-medieval period (Čulíková 2007, 2008).

Silene dichotoma was reported from medieval archaeobotanical samples in Uherský Brod (Opravil 1993).

Silybum marianum was used as a medicinal plant in the High Medieval (CZAD, Archaeological Institute ASCR 2011).

Sorbus domestica. Recorded from the Medieval repeatedly by Opravil (1994) and in the CZAD (Archaeological Institute ASCR 2011).

Vicia ervilia. Paleobotanical evidence suggests that the species was planted in the region in the Iron Age (Opravil 2000).

Another valuable source proved to be the summary of medieval sources on the use of medicinal plants in Bohemia (Čížek 1994). This author extracted information from the writings of Křišťan z Prachatic (Cristannus de Prachaticz, probably 1366–1431), a dean of the faculty of medicine and rector of Charles University in Prague, who wrote his works at the beginning of the 15th century, and they are probably the first scientific popularization in Czech. His Medicinal books and herbal, together with other then sources analysed by Čížek (1994), became, after careful interpretation of the original plant names, a basis for reclassifying the following species from neophytes to archaeophytes: Allium fistulosum, A. porrum, Angelica archangelica subsp. archangelica, Borago officinalis (see also Jankovská 2011, who gives archaeobotanical evidence from Prague and Opava in the High Medieval), Cnicus benedictus, Glycyrrhiza glabra, Hyssopus officinalis, Lactuca sativa, Lavandula angustifolia, Majorana hortensis, Ocimum basilicum, Paeonia officinalis, Pimpinella anisum, Ruta graveolens and Vicia faba.

Archaeophytes reclassified as neophytes

Lathyrus aphaca is considered alien without residence time specified in the national literature (Chrtková et al. 1977, Bělohlávková & Chrtková in Slavík 1995), and was classified as an archaeophyte in Pyšek et al. (2002). However, its status is reassessed here because the species is absent from old floras. Its earliest record for the country, based on a find in the vicinity of Uherské Hradiště, was published in 1856 (Šapetza 1856) and remained neglected, for instance, by Oborny (1886). The species only started to be occasionally recorded at the beginning of the 20th century (Chrtková et al. 1977).

Fumaria parviflora was reclassified based on reinterpretation of the account in the Flora of the Czech Republic (Smejkal in Hejný & Slavík 1988) and classification for Germany (Jäger 2011).
Species included in 2002 version and omitted from here

Seventy-five taxa listed in the previous version of the catalogue (Pyšek et al. 2002) were removed. They can be divided into several groups:

(i) Reclassified as native (41). For several taxa more or less convincing arguments were given recently suggesting their native status: Agropyron pectinatum (Repka & Chytrý in Hadinec et al. 2003), Crocus heuffelianus (Chrtk in Štěpáneková 2010), Epilobium dodonaei (Kaplan in Hadinec & Lustyk 2007), Senecio rupestris (Lustyk & Šída in Hadinec & Lustyk 2008), Teucrium scorodonia (Hadinec in Hadinec & Lustyk 2012) and Viola tricolor subsp. curtisi (V. Grulich, pers. comm.). This concerns mostly rare species occurring in a single or a few localities; the status of these species in the literature has been long debated without clear evidence in one direction or another. Following a conservative approach, we classify these species as native. Such a conservative approach also resulted in removing from the list some species originally classified as archaeophytes (Pyšek et al. 2002) yet without sufficient support for the hypothesis of their alien origin; they are thus considered native here: Aethusa cynapium, Androsace maxima, Arctium minus and its hybrid, A. xmaassii, Arnoiseris minima, Carduus crispus (including its hybrids C. ×stangii and C. ×sepincola), Cerinthe minor, Chenopodium ficifolium, C. glaucum, C. opulifolium, Cirsium vulgare (including its hybrids C. ×bipontinum, C. ×gerhardtii, C. ×sabaudum and C. ×subspinuligerum), Crepis biennis, Echium vulgare, Galeopsis ladanum, Medicago lupulina, Mentha arvensis (including its hybrids M. ×dalmatica and M. ×verticillata; see also Štěpánek 1998a, b), Pastinaca sativa subsp. sativa, Plantago major subsp. major (including its hybrids P. ×mixta and P. ×moravica), Polygonum aviculare, Sagina apetala subsp. apetala, S. apetala subsp. erecta, Scleranthus annuus and Vicia hirsuta.

In the case of Brachypodium rupestris, a possibility of its native status was discussed recently (Dančák & Hadinec in Hadinec & Lustyk 2011), but we consider it as a neophyte. This species has been reported as occurring in the territory of the current Czech Republic since the mid 19th century (Opiz 1852) albeit without unambiguous herbarium evidence; recently it has been discovered in several localities in eastern Moravia and near Veltěže in the České středohoří hills, northern Bohemia, as a clone growing on an area of approximately 7 m². The locality was destroyed in 1977 but the possibility of its native status was discussed recently (Dančák & Hadinec in Hadinec & Lustyk 2012). Although its native status is considered unlikely in the flora of the Czech Republic (Grulich in Slavík & Štěpánková 2004) and the species was listed as alien in Pyšek et al. (2002), the recent treatment suggests that its alien status be reconsidered (Hadinec & Lustyk 2012). We do not follow this opinion as the species is absent from historical floristic literature from this botanically very intensively studied area, and its native distribution is in southern Europe with the northernmost occurrences in Hungary, 400 km from the site in the Czech Republic (Grulich in Slavík & Štěpáneková 2004).

(ii) Not escaping from cultivation (9). Several taxa are not planted in the Czech Republic, or if they are, there is so far no evidence for them escaping from cultivation: Amelanchier ovalis (only rarely planted and not escaping; Lepší & Lepší 2008), Avena nuda (probably never cultivated, reports on its occurrence are confusing and relate to Avena sativa Chinesis Group; J. Zázvorka in Štěpáneková in prep.), Campanula speciosa (previous reports on escapes assigned to this species, native to the Pyrenees, most likely refer to C. glomerata), Catananche caerulea (listed previously based on a note about escape from cultivation in Dostál 1989 but herbarium evidence is lacking; Škalická in Slavík & Štěpánková 2004), Cerastium biebersteinii (this species, endemic to the Crimea, is most likely not planted in the Czech Republic, being confused with C. tomentosum), Cichorium intybus subsp. foliosum, Grindelia squarrosa, Ellisia nyctelea and Teucrium marum (no reliable records of escape from cultivation exist).

(iii) Taxonomically not justified taxa (10). This concerns some subspecies recognition of which is not justified based on the material from the Czech Republic; they are now included within the species level: Arrhenatherum elatius subsp. bulbosum, Bromus hordeaceus subsp. pseudothominii (included in B. hordeaceus subsp. hordeaceus; plants roughly corresponding to this taxon cannot be separated from other morphotypes of this highly variable species) and B. secalinus subsp. decipiens (included in B. commutatus). Further, this category includes some taxa with doubtful taxonomic status: Chenopodium integrifolium (included in Dysphania...
ambrosioides), *Hesperis matronalis* subsp. oblongipetala (included in *H. matronalis* subsp. *matronalis*), *Lathyrus articulatus* (included in *L. clymenum*), *Urtica dodartii* (included in *U. pilulifera*), *Vicia cordata* (included in *V. sativa*). Also excluded are some formerly listed hybrids: *Spergula arvensis* subsp. *arvensis* × *S. arvensis* subsp. *sativa*, and *Cannabis ×intersita* (a hybrid between two varieties, not distinguished in the current list).

(iv) Doubtful records (16). This category includes taxa that were recently suggested to have never occurred in the country, mistaken with other species, or bearing names that are difficult to interpret.

*Aster parviflorus* (syn. *Symphyotrichum parviflorum*). Reportedly an almost sterile taxon that originated in Europe but its possible occurrence and distribution in the Czech Republic is unclear (Kovanda & Kubát in Slavík & Štěpánková 2004).

*Bromus inermis* × *B. pumilianus*. The reexamination of plants growing at the locality reported in Krahulec & Jiříšte (1997) suggests that they fall within the range of individual variability of the native species *B. inermis* (B. Trávníček, pers. comm). Still, the issue may require further study.

*Bromus riparius*. A doubtful record without any details (Kubát et al. 2002) and unclear source.

*Bromus grossus*, recorded as *B. secalinus* subsp. *multiflorus* in Pyšek et al. (2002), is a weed of spelt wheat (*Triticum aestivum* Spelta Group) fields. It appears that this species was never documented from the Czech Republic as no specimens was found in Czech herbaria (J. Danihelka & J. Chrtek, unpubl.); the name was misapplied to plants of *B. secalinus* with spikelets consisting of many florets (see Dostál 1989).

*Centaurea nigra* × *C. phrygia* was listed based on the determination by the collector (V. Jehlík, PRA) but this hybrid combination has not been reported elsewhere and its occurrence is unlikely (P. Koutecký, pers. comm.).

*Cirsium ×reiseri*. Listed in previous version of this catalogue based on Dostál (1989) but not documented from the Czech Republic (Bures in Slavík & Štěpánková 2004).

×*Conygeron huelsenii*. This hybrid is reported in the literature since the 19th century (Čelakovský 1888b) but the herbarium specimens either represent different taxa, or are not available for some records. Although it is documented from neighbouring countries and its occurrence in the Czech Republic is possible, we omit it from the list due to the lack of evidence.

*Filago gallica*. Reported in Kubát et al. (2002) but the more recent treatment concluded that the occurrence of the species in the Czech Republic is doubtful and never reliably documented. One herbarium specimen available refers to *F. minima* (Štěch in Slavík & Štěpánková 2004).

*Filipendula rubra*. Reports on the occurrence of this species refer to *F. kamtschatica* (Slavík 2002).

*Hyacinthella ramelica* was reported by Šuk (2001) but not included in the recent treatment in the Flora of the Czech Republic, where Velká hora hill near Karštejn, central Bohemia, is mentioned as locality of *H. cf. leucophaea* (Bělohlávková in Štěpánková 2010). In fact, recent sources (Tutin et al. 1980, Delipavlov et al. 2003) recognize only *H. leucophaea*, even without mentioning *H. ramelica* in its synonymy. In our treatment, the plants reported by Šuk (2001) as *H. ramelica* are therefore included within *H. leucophaea*.

*Kickxia elatine* subsp. *crinita*. Chrtek (1984), analysing the variation of the populations of *K. elatine* subsp. *elatine* in southern Moravia, considered some of the morphotypes transitory towards this Mediterranean subspecies. He even identified one specimen as *K. elatine* subsp. *crinita* (see also Slavík in Slavík 2000). However, based on phytogeographic information, we consider its occurrence in the Czech Republic quite unlikely and include all records of *K. elatine* in the type subspecies.

*Lithospermum arvense* subsp. *caeruleascens* is reported to have occurred near Všetaty, central Bohemia (Slavík in Slavík 2000). Given that both *Buglossoides arvensis* and *B. incrassata* have blue-flowered forms, it is impossible to interpret the above report with certainty.

*Mantisalca salmantica*. Rather vague literature reports about occasional occurrence of this species in the Czech Republic (Dostál 1989, Kubát et al. 2002) lack details and are not supported by herbarium specimens (Štěpánek in Slavík & Štěpánková 2004).

*Parapholis strigosae*. The species was reported by Dostál (1989: 1357) as “once introduced to Brno with cotton”. We believe that this is a misinterpretation based on the record of *Pholiurus incurvus (= Parapholis incurva)*, introduced to Brno with wool and reported earlier by Dvořák & Kühn (1966). No specimens documenting the occurrence of *P. strigosae* in the Czech Republic were found in herbaria.

*Veronica acinifolia*. The only herbarium specimen from the Czech Republic, on which the reported occurrence is based (Smejkal 1970, Hrouda in Slavík 2000), belongs to *V. triphylos* (Danihelka 2011).

*Vicia ×smejklavka*. Omitted due to the lack of evidence.
Pending issues: species with uncertain status, doubtful records and taxa requiring further study or monitoring

In the Průhonice Park near Prague there is a good long-term record of regeneration of planted woody taxa. The following were observed to regenerate, mostly from seed, in the vicinity of planted individuals: *Acer saccharum*, *Carya ovata*, *C. tomentosa*, *Crataegus intricata*, *Fraxinus rhynchophylla*, *Chamaecyparis nootkatensis*, *C. pisifera*, *Juglans ailanthifolia*, *Liriodendron tulipifera*, ×*Malosorbus florentina*, *Mahonia repens*, *Padus maackii*, *Picea sitchensis*, *Pterocarya stenoptera*, *Quercus palustris*, *Rhododendron luteum*, *Symlocos paniculata*, *Taxus baccata* × *T. cuspidata*, *T. cuspidata*, *Thuja plicata*, *Torreya nucifera* and *Tsuga canadensis* (J. Burda, pers. comm.). These taxa are not included in the list but a note is given here for comparison with other regions of the world where they may appear as aliens.

Some species are not included in the list even though they are reported in national sources such as floras and field guides. This concerns, for example, several taxa of the genus *Symphyotrichum* (*Aster* s. l.), possible occurrence and distribution of which in the Czech Republic is unclear and requires further study. These species are either reported as being confused with other species in older sources (*S. tradescantii*), or are reported as (likely) to occur but not documented by any herbarium specimens (*S. praecatum*, *S. ericoides*). This is also the case of *Galatella sedifolia* subsp. *sedifolia* (syn. *Aster punctatus*). Two records exist from the Czech Republic, both assuming either accidental introduction or garden escape (Makowsky in Oborny 1885, see also Danihelka 2008, Dostál et al. 1948–1950) but no herbarium specimens were found.

The same conservative approach was adopted towards hybrids with alien species involved in taxonomically difficult genera that are reported from the Czech Republic but not confirmed with certainty, e.g. *Chenopodium album* × *C. strictum*, *C. × tridentium* (= *C. opulifolium* × *C. strictum*), *C. × variabile* (= *C. album* × *C. berlandieri* subsp. *zschackei*; Dostálek et al. in Hejný & Slavík 1990, Kubát et al. 2002), or *Atriplex hortensis* × *A. sagittata* (Kubát et al. 2002).

We did not include species that are known to have been planted in the wild and survive as the originally planted individuals such as *Rhododendron hirsutum* and *R. ferrugineum* (Kubát et al. 2002). One shrub of the latter species was planted in the Králický Sněžník Mts and still survives since at least 1825 when it was first recorded (F. Krahulec, pers. obs.). Neither were included cases such as *Cyclamen coum*, of which one plant was found in the Radotínské údolí valley, Prague, where it was most likely deliberately planted and reported to survive since 2008 (Prančl in Hadinec & Lustyk 2012).
Appendix 2. – List of alien taxa of the Czech flora. Taxa are arranged alphabetically. **Family** codes (Fam) are formed by initial letters of the family name. The following information is given for each taxon, if available: **Life history** (LH): a – annual, b – biennial, pe – perennial, ss – semishrub, s – shrub, t – tree, f – fern, aq – aquatic, p – parasitic (life histories in which the taxon does not occur in the Czech Republic are given in parentheses). **Residence time status** (Res): ar = archaeophyte, neo = neophyte. **Invasion status** (Inv): cas = casual, nat = naturalized, inv = invasive. **Population group** (PG): 1–18, reflecting the dynamics of populations of the species in the region, with link to cultivation (see text for details). **First record** (1st): date of the first reported occurrence in the wild in the Czech Republic; in some cases approximate date (century or decade) is given inferred from the sources (e.g. 17th, 1990s). **Abundance type** (Abund) in the wild in the country: s – single locality, r – rare, sc – scattered, la – locally abundant, c – common, v – vanished (if no records have been known for a long period), s+ev – single locality, now vanished. **Pathway of introduction** (Path) of the species into the country: d – deliberate planting involved; a – accidental (unintentional) pathway only. **Region of origin**: M – Mediterranean region, E – Europe, As – Asia, Af – Africa, AmN – North America, AmC – Central America, AmS – South America, Au – Australia, hybrid – hybrid origin, anec – anecophyte (see text for details). **Origin** includes the hybrid formulas for hybrids (nothospecies) and most anecophytes of hybrid origin listed here under their binomials are given in Electronic Appendix 2. **Cover** refers to average % cover in plant communities in the Czech Republic; upper index refers to the number of vegetation plots from which the value was calculated (note that the same values are given for *Chenopodium striatiforme* and *C. strictum*, and *Prunus domestica* and *P. insititia*, respectively, as the vegetation plots with these species could not be distinguished with certainty, and were merged). **Number of habitats** (Hab), classified according to Sádlo et al. (2007), in which the species grows (n = 88). **Impact** (IEc – ecological, IEn – economic): yes indicates that the species is reported to exert an impact in Europe; yes+, documented from the Czech Republic. **Source**: It primarily refers to the treatment in the Flora of the Czech Republic if the species is reported there as an alien; otherwise the sources refer to papers first reporting the species, or explicitly dealing with the given taxon. Also included are selected comprehensive accounts and specialized case studies, or updates of recent situation. Detailed information on taxa that represent additions to the Czech flora is given in Appendix 1. References to the eight volumes of the Flora of the Czech Republic (F1 – Hejný & Slavík 1988, F2 – Hejný & Slavík 1990, F3 – Hejný & Slavík 1992, F4 – Slavík 1995, F5 – Slavík 1997a, F6 – Slavík 2000, F7 – Slavík & Štěpánková 2004, F8 – Štěpánková 2010) and to the Additamentum ad floram Reipublicae Bohemicae series (A1 – Hadinec et al. 2002, A2 – Hadinec et al. 2003, A3 – Hadinec et al. 2004, A4 – Hadinec et al. 2005, A5 – Hadinec & Lustýk 2006, A6 – Hadinec & Lustýk 2007, A7 – Hadinec & Lustýk 2008, A8 – Hadinec & Lustýk 2009, A9 – Hadinec & Lustýk 2011) are indicated using codes. Taxa reported for the first time here are designated as ‘this study’. See Appendix 1 for comments on newly added and/or taxonomically difficult taxa, and for changes in residence time status.

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<th>Taxon</th>
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<td><em>Chaenomeles japonica</em> (Thunb.) Spach</td>
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<td><em>Chaerophyllum nodosum</em> (L.) Crantz</td>
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<td>Filipov 1999</td>
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<td>Fabt s neo cas 9 r del E</td>
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<td><em>Chelidonium majus</em> L.</td>
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<td>Smejkal 1952, Štěpánek &amp; Holub in F5</td>
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<td><em>Cephalaria syriaca</em> (L.) Roem. et Schult.</td>
<td>Dips a neo cas 1 r acc M 4</td>
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<td><em>Cerastium arvense</em> subsp. arvense × C. tomentosum</td>
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<td>Smejkal in F2</td>
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<td><em>Chenopodium berlandieri</em> subsp. zschackei (Murr) Zobel</td>
<td>Amara a neo cas 1 r acc AmN 4&lt;sup&gt;68&lt;/sup&gt; 9</td>
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<td>Amara a neo cas 3 1809 r del AmN</td>
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 geral: catalog of alien plants of the Czech Republic, Pyšek et al. (2002)
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This table includes information on the distribution, origin, and habitat of various alien plant species in the Czech Republic.
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**Note:** The table includes columns for Family (Fam), Life History (LH), Residence (Res), Invasion (Inv), Plant Group (PG), 1st Abundance (1st), Pathogenicity (Path), Origin, Cover, Habit (Hab), Ecological (IEc), Endemic (IEn), and Source.
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<td>a</td>
<td>ar</td>
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